



**HAL**  
open science

## **Grazing intensity modulates carbohydrate storage pattern in five grass species from temperate grasslands**

Marie-Lise Benot, Annette Morvan-Bertrand, Cendrine Mony, Julia Huet, Cécile Sulmon, Marie-Laure Decau, Marie-Pascale Prud'Homme, Anne Bonis

### ► **To cite this version:**

Marie-Lise Benot, Annette Morvan-Bertrand, Cendrine Mony, Julia Huet, Cécile Sulmon, et al.. Grazing intensity modulates carbohydrate storage pattern in five grass species from temperate grasslands. *Acta Oecologica*, 2019, 95, pp.108-115. 10.1016/j.actao.2018.11.005 . hal-01983581

**HAL Id: hal-01983581**

**<https://univ-rennes.hal.science/hal-01983581>**

Submitted on 16 Jan 2019

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

**Grazing intensity modulates carbohydrate storage pattern in five grass species from temperate grasslands.**

Marie-Lise Benot<sup>1,2,\*</sup>, Annette Morvan-Bertrand<sup>3</sup>, Cendrine Mony<sup>1</sup>, Julia Huet<sup>1</sup>, Cécile Sulmon<sup>1</sup>, Marie-Laure Decau<sup>3</sup>, Marie-Pascale Prud'homme<sup>3</sup>, Anne Bonis<sup>1,4</sup>

<sup>1</sup>Univ Rennes, CNRS, ECOBIO [Ecosystèmes, biodiversité, évolution] - UMR 6553, F-35000 Rennes, France

<sup>2</sup>BIOGECO, INRA, Univ. Bordeaux, F-33615 Pessac cedex, France

<sup>3</sup>Ecophysiologie Végétale Agronomie et nutritions N.C.S. Normandie Univ, UNICAEN, INRA, EVA, 14000 Caen, France

<sup>4</sup>Present address: UMR GEOLAB, CNRS-Université Clermont Auvergne, 4 rue Ledru, 63057 Clermont-Ferrand cedex 1, France

\*Corresponding author. Email: [marie-lise.benot@u-bordeaux.fr](mailto:marie-lise.benot@u-bordeaux.fr).

**Running head.** Effect of grazing intensity on carbohydrate storage.

**Declarations of interest:** none.

1 **Abstract**

2 Regrowth after defoliation is an essential mechanism of plant tolerance to grazing. In grasses,  
3 non-structural carbohydrates (NSC) contained in tiller bases constitute a major substrate for  
4 regrowth after defoliation. Using a multi-specific approach, the present study aimed at testing  
5 the effect of grazing intensity on NSC concentration in tiller bases. We selected five grazing-  
6 tolerant grass species (*Agrostis stolonifera*, *Cynosurus cristatus*, *Hordeum secalinum*, *Lolium*  
7 *perenne* and *Poa trivialis*) and collected plants in a grassland subjected to two cattle grazing  
8 intensities (intensive *versus* moderate) for years. We measured NSC concentrations (starch,  
9 fructans, sucrose, glucose and fructose) in tiller bases. We found that fructan and sucrose  
10 concentrations before the grazing season (April) were higher under intensive than moderate  
11 grazing. By contrast, no significant effect of the grazing intensity on these NSC  
12 concentrations in tiller bases remained at the end of the grazing season (October). These  
13 results suggest that the level of reserves available before the onset of disturbance caused by  
14 grazing as well as the reserve replenishment capacity during the grazing season are modified  
15 by the intensity of grazing.

16

17 **Key-words:** Fructans, grassland, HPLC, non-structural carbohydrates, tolerance to grazing,  
18 tiller bases.

19

20 **Abbreviations:** DM: dry mass, HPLC: high performance liquid chromatography, NSC: non-  
21 structural carbohydrates, WSC: water-soluble carbohydrates

22

23

24

25

## 26 **Introduction**

27 Grazing acts as a complex factor on vegetation, inducing major changes in floristic and  
28 functional composition in grassland plant communities (Bullock *et al.*, 2001; Diaz *et al.*,  
29 2001, 2007; de Bello *et al.*, 2005; Marion *et al.*, 2010). It drives vegetation changes mainly  
30 through the consumption of above-ground biomass *i.e.*, defoliation (Kohler *et al.*, 2004).  
31 Defoliation affects individual plants directly, by removing part of photosynthetically active  
32 biomass (leaves and stems) or indirectly, through canopy opening and subsequent decrease in  
33 aboveground competition (Sala *et al.*, 1986; Richards, 1993; Bakker *et al.*, 2003; Bonis *et al.*,  
34 2005; Veen *et al.*, 2008). Plant ability to survive and develop after defoliation depends on  
35 compensatory growth, mainly based on the plant ability to regrow after damage  
36 (McNaughton, 1983; Maschinski and Whitham, 1989), an essential mechanism of tolerance to  
37 grazing (Briske, 1996; Stowe *et al.*, 2000). Considering grazing as a filter over the local  
38 species pool, it generally favours plants capable of rapid regrowth after defoliation enabling  
39 the recovery of leaf area and the resumption of photosynthetic activity (Fulkerson and  
40 Donaghy, 2001).

41 Regarding the source of carbon, leaf regrowth after defoliation generally takes place into two  
42 phases. First, defoliation is followed by a transient time lag, during which stored  
43 carbohydrates are mobilized, a major mechanism that enables the plant to recover from losses  
44 of photosynthetic tissues (Richards, 1993; Morvan-Bertrand *et al.*, 1999a; Morvan-Bertrand *et*  
45 *al.*, 1999b; Schnyder and de Visser, 1999). Second, photosynthetic activity of newly produced  
46 tissues becomes the main source of assimilates, supporting growth and replenishment of  
47 reserve pools (Richards, 1993; de Visser *et al.*, 1997; Morvan-Bertrand *et al.*, 1999b).  
48 Consequently, tolerance to grazing, which depends on enhanced short-term ability of leaf area  
49 recovery after grazing, is likely to mainly rely on easily mobilized reserves (Richards, 1993).

50 Grazing occurs within a range of intensity: even in semi-natural grasslands the grazing  
51 pressure may vary from extensive to intensive, due to management choice but also to local  
52 heterogeneity in the animal choice (Marion *et al.*, 2010; Loucougaray *et al.*, 2004). Intensive  
53 grazing may corresponds to severe and/or frequent defoliation events, with respectively large  
54 amounts of biomass removed by each defoliation event and short time lags for reserve  
55 replenishment. Frequency between two consecutive defoliation events is a dimension of  
56 grazing intensity by which reserve pools may be substantially depleted (Beaulieu *et al.*, 1997;  
57 Kleijn *et al.*, 2005). Reserve making is considered to be costly, as it diverts resources from  
58 potential growth, decreasing growth rate and leading to smaller and less competitive plants  
59 (Kobe, 1997; van der Meijden *et al.*, 2000; de Jong and van der Meijden, 2000). Thus, reserve  
60 making and remobilization for compensatory growth are expected to be selected for in  
61 environments where their benefits outweigh these costs (Karban and Baldwin, 1997). In  
62 particular, the capacities to constitute large pools of reserves available before the beginning of  
63 a grazing season and to quickly remobilize and replenish these pools after defoliation (Lee *et*  
64 *al.*, 2010) are both expected to enable plants to cope with intensive grazing.

65 Storage of resources (carbon and other elements such as nitrogen) can take place in a diversity  
66 of plant organs such as roots, perennating organs and stems (van der Meijden *et al.*, 1988;  
67 Suzuki and Hutchings, 1997; Klimeš and Klimešová, 2002; Kavanová and Gloser, 2005), but  
68 considering a wide range of species, the highest levels are found in leaves (Martínez-Vilalta *et*  
69 *al.*, 2016). In particular, grasses store large amounts of carbohydrates in tiller bases, which is  
70 composed of elongating leaves enclosed in sheaths of mature leaves (Morvan-Bertrand *et al.*,  
71 1999a; Morvan-Bertrand *et al.*, 1999b). Carbohydrates contained in tiller bases are mobilized  
72 rapidly after defoliation and, due to their close proximity to the leaf growth zone, they are  
73 likely to play a key role in compensatory growth (Morvan-Bertrand *et al.*, 2001). Carbon  
74 reserves, mainly stored as non-structural carbohydrates (NSCs), have indeed been shown to

75 be of great importance in regrowth after defoliation. Starch constitutes the main form of long-  
76 term storage NSCs in many plant species (Manner, 1985). In grasses, starch is the major  
77 carbohydrate stored in seeds. By contrast, fructans (polymers of fructose) are the predominant  
78 compounds of carbon storage in vegetative parts in most of C3 grasses from temperate areas  
79 (Pollock and Cairns, 1991). As the main extracellular form of NSCs, sucrose (non-reactive  
80 disaccharide) is involved in source-sink carbon transport (Salerno and Curatti, 2003) and  
81 might act as a signalling compound in response to environmental cues (Hawker, 1985; Rosa *et*  
82 *al.*, 2009). Sucrose represents the dominant form of carbohydrate storage (Hawker, 1985) in  
83 some species, such as sugar beet (*Beta vulgaris*, Chenopodiaceae) or sugar cane (*Saccharum*  
84 *sp.*, Poaceae), and can also constitute a major form of carbon reserves in vegetative tissues of  
85 temperate C3 grasses beside fructans (Chatterton *et al.*, 1989).

86 The content and composition of carbon reserves can vary at several temporal scales from the  
87 day to the year. In temperate climates with seasonal variations, numerous studies reported  
88 annual dynamics, with the highest pools of reserves recorded in late summer or fall (e.g.,  
89 Pollock and Cairns, 1991; Beaulieu *et al.*, 1997; Kleijn *et al.*, 2005; Asaeda *et al.*, 2006;  
90 Janeček *et al.*, 2011; Baptist *et al.*, 2013; Benot *et al.*, 2013b; Janeček *et al.*, 2015), after a  
91 period of slow growth or leaf senescence and nutrient recycling (Chapin *et al.*, 1990). Then,  
92 reserve pools are generally reported to decrease during winter and during the early phases of  
93 the growing period in spring, certainly due to the combination of reduced photosynthesis and  
94 of reserve remobilisation to support the maintenance of metabolism during winter and the  
95 spring regrowth outbreak (Bloom *et al.*, 1985; Fulkerson and Donaghy, 2001). In most  
96 grasslands from temperate climate and, more particularly in commonly grazed grasslands,  
97 which are widespread along the Atlantic coast, the grazing season generally expands from  
98 spring to autumn i.e., during the period of resource storage in plants. The seasonal processes  
99 of resource storage are thus likely to be altered by grazing.

100 The present study aimed at studying how the pattern of formation and replenishment of  
101 carbohydrate reserves in tiller bases is modulated by the grazing intensity. More precisely, we  
102 formulated the two following hypotheses. H1 (hypothesis of efficient formation of reserves):  
103 plant species tolerance to grazing is expected to rely on large pools of reserves available  
104 before the beginning of a grazing season. Thus, for plant species tolerant to grazing, the  
105 concentrations of carbohydrate reserves before the beginning of a grazing season are expected  
106 to be higher under intensive than under moderate grazing. H2 (hypothesis of efficient  
107 replenishment of reserves): plant species tolerance to grazing is expected to rely on efficient  
108 reserve replenishment between two consecutive defoliation events. Thus, regardless of the  
109 grazing intensity, plant species tolerant to grazing are expected to be able to quickly restore  
110 reserve pools, and to show equivalent concentrations of carbohydrate reserves at the end of  
111 the grazing season whatever the grazing intensity.

112 We tested these hypotheses for five Poaceae species from a grassland traditionally grazed for  
113 6 months a year from early spring (April) to early fall (October), where cattle grazing has  
114 generated mosaics of vegetation patches (Marion *et al.*, 2010). Plants were collected in  
115 vegetation patches submitted to two contrasting cattle grazing intensities (moderate and  
116 intensive) for about 15 years, at two sampling dates: just after a 6-month grazing season  
117 (October) and just before the beginning of the following grazing season (April). NSC  
118 concentrations (starch, fructans, sucrose, glucose and fructose) were measured in tiller bases  
119 of individual plants.

120

121

122

123

124

## 125 **Material and methods**

126

### 127 **Study site**

128 This study was carried out on the most abundant species in the vegetation in a commonly-own  
129 grassland of Magnils–Reigniers (250 ha-large), in Marais poitevin (French Atlantic coast,  
130 46° 28'N; 1° 13'W). This grassland has been reclaimed from the sea from the X<sup>th</sup> century  
131 onwards. A topographical gradient consisting of depressions, intermediate slopes and higher-  
132 level flats (with a maximum altitudinal range of 0.7 m) occurs repeatedly within the grassland  
133 (Amiaud *et al.*, 1998) and discriminates three plant communities, depending on flooding  
134 duration: the hygrophilous, meso-hygrophilous and mesophilous communities, respectively  
135 (Marion *et al.*, 2010). As fructans have been suggested to promote plant resistance to hypoxia  
136 (Albrecht *et al.*, 1997), this study was conducted on species collected within the never-  
137 flooded mesophilous community, to avoid confounding effects of the flooding regime on  
138 fructan concentrations. The soil on flats are sodisols characteristic of pasture grasslands in the  
139 Marais poitevin (Amiaud *et al.*, 1998).

140 This grassland has traditionally been grazed by cattle and horses from April to October, a  
141 grazing season occurring during spring and summer (Amiaud *et al.*, 1998). An experimental  
142 design consisting in several paddocks has been set up in this grassland and has enabled to  
143 control the grazing intensity (from no to intensive grazing) and herbivore type (cattle and/or  
144 horses) since 1995 (Loucougaray *et al.*, 2004; Rossignol *et al.*, 2006). The present study only  
145 focused on vegetation grazed by cattle. Since 1995, cattle grazing has generated mosaics of  
146 vegetation patches submitted to contrasting grazing intensities, with stabilised composition  
147 and structure (Marion *et al.*, 2010). The study species were collected in moderately (G+) and  
148 intensively (G++) grazed patches within 1ha-large paddocks grazed by 2 heifers.ha<sup>-1</sup> (*i.e.*  
149 about 685 kg ha<sup>-1</sup>) and 4 heifers.ha<sup>-1</sup> (*i.e.* about 1370 kg ha<sup>-1</sup>, Ménard *et al.*, 2002) respectively



150 (see also Benot *et al.*, 2013a). Moderately grazed vegetation (G+) is characterised by diverse  
151 grass species such as *Elytrigia repens* (L.) Nevski, *Cynosurus cristatus* L. and *Hordeum*  
152 *secalinum* Schreb, whereas intensively grazed vegetation (G++) is dominated by the grass  
153 *Lolium perenne* L. (Marion *et al.*, 2010).

154

#### 155 **Study species**

156 We selected five common perennial Poaceae species of the mesophilous plant community that  
157 are present under both moderate and intensive grazing (Table 1). *Agrostis stolonifera* L.,  
158 *Cynosurus cristatus* L., *Hordeum secalinum* Schreb., *Lolium perenne* L. and *Poa trivialis* L.  
159 are forage species, frequently found in meadows and pastures (Grime *et al.*, 2007), with a  
160 high level of grazing tolerance (see Table 1 for species level of tolerance to grazing retrieved  
161 from the TRY database, Kattge *et al.*, 2011; original source: BiolFlor database, Kühn *et al.*,  
162 2004). They are all tussock-forming and do not present organs potentially specialized in  
163 storage (e.g. tubers, rhizomes). *A. stolonifera* generally produces long creeping stems  
164 (stolons) but displays an important morphological plasticity. This species was included in the  
165 study as plants growing in the mesophilous community remain mainly tussock forming with  
166 only a few or short stolons (personal observations). Thus tiller bases were considered as the  
167 main storage organ for the five study species.

168 Although present under both grazing intensities, study species vary in abundance in the  
169 vegetation. In June 2008 *i.e.*, during the biomass peak, we estimated these variations by  
170 recording the relative cover of the five species in ten 0.5 m × 0.5 m plots randomly located  
171 within the mesophilous community in G+ and G++ vegetation patches (Table 1). Given the  
172 co-occurrence of the study species at very small scales (*i.e.*, even less than 0.25 m<sup>2</sup>, personal  
173 observations), these variations in abundance are very little likely to be caused by selective  
174 cattle grazing. On the contrary, leaf damages caused by cattle defoliation were commonly

175 observed for all of these species (personal observations) and variations in species abundance  
176 likely reflect differences in tolerance to grazing.

177

### 178 **Plant collection and material**

179 Eight non-flowering individuals of each species, composed of several connected tillers, were  
180 randomly collected within the mesophilous community under both grazing intensities (G+ and  
181 G++ patches), just after a 6-month grazing season (October 2008) and two days before the  
182 following grazing season (April 2009). In October, plants were collected about ten days after  
183 the heifers had left the pastures and showed traces of leaf damage, indicating former grazing.  
184 This time lag from the end of the grazing season was chosen to ensure that the phase of  
185 reserve remobilization was completed and reserve replenishment could have been initiated  
186 (Richards, 1993; Morvan-Bertrand *et al.*, 1999a). In contrast, plants collected in April were  
187 intact. At both dates, only non-flowering tillers were selected. Tiller collection lasted for two  
188 consecutive days, from 10am to 4pm. We paid attention to pick up tillers of all species all day  
189 long, in order to capture the daily variation of carbohydrate content within a species (Chapin  
190 *et al.*, 1990; Shewmaker *et al.*, 2006; Smith and Stitt, 2007; Haydon *et al.*, 2011). Each  
191 sample consisting in a few connected tillers was picked up with a knife, carefully washed and  
192 immediately frozen in liquid nitrogen. From the field to the lab, samples were transported in a  
193 freezer and kept at -80°C until they were freeze-dried. For each sample, a tiller was randomly  
194 selected and freeze-dried plants were dissected in order to separate tiller basis, composed of  
195 mature leaf sheaths and enclosed elongating leaves, from the rest of the tiller (*i.e.*, leaf blades  
196 and roots). As fructans accumulate mainly in the first centimeters of the tiller bases (Pollock  
197 and Cairns, 1991; Morvan-Bertrand *et al.*, 2001) and as tillers might differ in size, we  
198 considered NSC concentrations in 3-cm tiller bases as a relevant indicator of the amount of  
199 carbon resources stored and potentially available for biomass compensation per biomass unit.

200 The tiller basis was thus cut either at 3 cm above the rooting point or under the ligule of the  
201 older leaf for tiller bases shorter than 3 cm. Tiller bases were then marble-powdered.

202

### 203 **Extraction, purification and separation of water-soluble carbohydrates (WSC)**

204 Twenty  $\pm$  1 mg dry mass (DM) of powder were weighted. Water soluble carbohydrates (WSC:  
205 fructans, sucrose, glucose and fructose) were extracted from this powder in 80% ethanol at  
206 80°C for 15 min. After ethanol extraction, the sample was centrifuged at 10,000 g for 10 min.  
207 The supernatant was preserved on ice and 2 mL of water were added to the pellet. The tube  
208 contents were mixed and incubated 15 min at 60°C. After the first aqueous extraction, the  
209 sample was centrifuged at 10,000 g for 10 min. The supernatant was preserved on ice and the  
210 aqueous extraction was repeated once with the pellet. The three supernatants were pooled,  
211 evaporated to dryness under vacuum and the residue was dissolved into 450  $\mu$ L of ultra-pure  
212 water. Aliquots of WSC extract (100  $\mu$ L) were passed through minicolumns (Mobicols from  
213 MoBITec, Göttingen, Germany) containing 150  $\mu$ L of anion exchange resin (Amberlite CG-  
214 400 II, formate form, Fluka, Buchs, Switzerland) and 250  $\mu$ L of cation exchange resin  
215 (Dowex 50W X8-400, H<sup>+</sup> form, Sigma, Saint-Louis, MO, USA) to remove charged  
216 compounds. Between these two resins, 80  $\mu$ L of PVPP (polyvinylpolypyrrolidone) were  
217 added to eliminate lipids, pigments and phenolic compounds.

218 Glucose, fructose, sucrose, and fructans were separated and quantified by high-performance  
219 liquid chromatography (HPLC). The eventual remaining impurities were removed by a pre-  
220 column Guard-PAK (Millipore Waters, Milford, MA, USA) and the WSC were then separated  
221 on a cation exchange column (Sugar-PAK I, 300  $\times$  6.5 mm, Millipore Waters Milford, MA,  
222 USA) eluted at 0.5 mL.min<sup>-1</sup> with 0.1 mM CaEDTA at 85°C, and detected using a  
223 refractometer as a sugar detector (see also Supplementary Material in Benot *et al.*, 2013b).

224 The concentration of each NSC was calculated as the carbohydrate mass divided by the dry  
225 mass of tissue powder.

226

### 227 **Starch measurement**

228 Starch insoluble pellet remaining after WSC extraction was dissolved under agitation into  
229 200  $\mu\text{L}$  of dimethylsulfoxide (DMSO) and 50  $\mu\text{L}$  of HCl (8N) at 60°C for 30 min. After  
230 centrifugation (10,000 g for 15 min), 200  $\mu\text{L}$  of extract were dissolved into 500  $\mu\text{L}$  of ultra-  
231 pure water added with 40  $\mu\text{L}$  of NaOH (5M) and pH was adjusted to 4.5. Ultra-pure water  
232 was then added up to 1mL. After decantation, starch content was measured using enzymatic  
233 kits (Enzyplus® kit EZ0 942+ Starch, Raisio Diagnostics SpA, Rome, Italy). In short, starch  
234 was degraded by amyloglucosidase into glucose, which was quantified through NADPH  
235 production by spectrophotometry at 340 nm (Sulmon *et al.*, 2011). Starch concentration was  
236 calculated as its mass divided by the dry mass of tissue powder.

237

### 238 **Statistical analyses**

239 Variation in the concentration of each NSC was analysed by linear model ANOVAs. First, in  
240 order to check for seasonal patterns of NSC concentrations in tiller bases, and to test whether  
241 grazing intensity modified these patterns, a model was carried out with the date, species and  
242 grazing intensity as main effects. Then, in order to test for the effect of grazing intensity on  
243 NSC concentrations just after and just before the beginning of a grazing season, models with  
244 species and grazing intensity as main effects were carried out separately for October and  
245 April. Tukey HSD tests were used for post-hoc comparisons. When necessary, data were  
246 either log-transformed or arcsine square root-transformed to improve homoscedasticity and  
247 normality of the residuals. In cases where extreme values in the dataset limited the effect of  
248 transformations, these values were temporally removed and the new models were compared

249 with complete ones to evaluate the robustness of the results. Only complete models based on  
250 all data were kept for the study. Because of errors in laboratory measurements, a few  
251 replicates were lost (see Table S1 for final number of replicates per date  $\times$  species  $\times$  grazing  
252 intensity interactions). Statistical analyses were carried out with the R software version 3.1.1  
253 (R Development Core Team, 2014, <http://www.R-project.org>) and post-hoc HSD test with the  
254 agricolae package (de Mendiburu, 2014).

255

## 256 **Results**

257 A general trend for higher NSC concentrations in October than in April was observed.  
258 However, this trend depended of the species and was affected by the grazing intensity (Table  
259 2, Figure 1). While the effect of the sampling date was significant for all of the five NSCs,  
260 significant date  $\times$  species interaction was recorded for all NSC but fructose ( $F_{1,133} = 2.33$ ,  $P =$   
261  $0.06$ ), a date  $\times$  grazing intensity interaction for all NSC but glucose ( $F_{1,133} = 3.29$ ,  $P = 0.07$ )  
262 and even a significant date  $\times$  species  $\times$  grazing intensity for fructose ( $F_{4,133} = 2.69$ ,  $P = 0.03$ ,  
263 see Table 2 for complete results of the model). Starch was the least abundant carbohydrate  
264 with concentrations in tiller bases lower than  $3 \text{ mg g DM}^{-1}$  in October and even than  $1 \text{ mg}$   
265  $\text{g DM}^{-1}$  in April (Figure 1A, F). Fructan concentrations in tiller bases were largely higher than  
266 that of the other NSCs, comprised on average between  $80$  and  $400 \text{ mg g DM}^{-1}$  depending on  
267 the species and the date (Figure 1B, G).

268 At the end of the grazing season (October), species effect was significant for all NSC  
269 concentrations (Figure 1A-E). For fructans, the highest concentrations were found in  
270 *L. perenne* (Figure 1B), which is the only species being much more abundant under intensive  
271 than moderate grazing (Table 1). *C. cristatus* showed the lowest fructan concentrations  
272 (Figure 1B). At this date, and considering the five species altogether, grazing intensity only

273 affected fructose concentrations, with generally larger concentrations under intensive grazing  
274 (Figure 1E).

275 Just before the following grazing season (April), a significant species effect for all NSCs  
276 except glucose was detected (Figure 1F-J). Fructan and sucrose concentrations were  
277 significantly affected by grazing intensity with higher values for plants submitted to intensive  
278 grazing (Figure 1G, H). In contrast, starch concentration was globally negatively affected by  
279 grazing intensity, despite a species-dependant effect. Considering hexoses, glucose  
280 concentration was not affected by the grazing intensity (Figure 1I), whereas grazing effect on  
281 fructose concentration depended on the species (Figure 1J).

282

## 283 **Discussion**

284 The ranks of NCS concentrations remained similar among the species investigated: fructans  
285 were the most abundant NSCs, followed by sucrose, glucose and fructose, while starch was  
286 only found as traces. The highly different concentrations of fructans and starch found in the  
287 present study confirmed the marginal role of starch as a reserve NSC in vegetative organs for  
288 most C3 grasses from temperate climates (Brocklebank and Hendry, 1989; Pollock and  
289 Cairns, 1991; Cairns *et al.*, 2002). Given their largely higher abundance in the study species,  
290 fructans and even sucrose emerged as the most relevant NSC in carbon storage.

291

### 292 **Seasonal variations in NCS concentrations**

293 We found that concentrations of reserve carbohydrates were globally higher at the end of  
294 summer than in early spring, providing further support to a commonly observed seasonal  
295 pattern (e.g., Steen and Larsson, 1986; Pollock and Cairns, 1991; Beaulieu *et al.*, 1997;  
296 Klimešová and Klimeš, 2003; Janeček *et al.*, 2011; Baptist *et al.*, 2013). However, this date  
297 effect depended on the species as well as on the grazing intensity, indicating that grazing

298 intensity modified the seasonal patterns of reserve formation. This seasonal variation in NSC  
299 concentrations can be explained by a higher rate of mobilization than of photosynthetic  
300 assimilation during winter and early spring. During winter, carbohydrate storage is likely to  
301 decrease because of lower photosynthetic activity due to shorter day length and lower  
302 incoming radiation and temperature (Fulkerson and Donaghy, 2001; Höglind *et al.*, 2011).  
303 Carbohydrate reserves can also be depleted during winter because reserves may sustain  
304 energy demand during unfavourable conditions (Bloom *et al.*, 1985). In addition, reserves  
305 may be depleted by early-spring consumption before the grazing season in order to support  
306 fast growth and tiller emergence (Chapin *et al.*, 1990; Pollock and Cairns, 1991; Beaulieu *et*  
307 *al.*, 1997; Kleijn *et al.*, 2005; Asaeda *et al.*, 2006). Consequently, the decrease observed in the  
308 present study in NSC concentrations and more particularly, in starch, fructan and even sucrose  
309 concentrations, between October and April may be due both to reduced production and  
310 remobilization.

311

### 312 **Just before the grazing season: a prevailing effect of grazing intensity**

313 April sampling was conducted to reflect carry-over effects of former grazing seasons. We  
314 detected no significant effect of the grazing intensity (moderate *versus* intensive grazing)  
315 formerly applied on the vegetation on starch concentration for three species (*C. cristatus*,  
316 *L. perenne* and *P. trivialis*) and this concentration was lower under intensive than under  
317 moderate grazing for two species (*A. stolonifera* and *H. secalinum*). By contrast, regardless of  
318 the species, fructan and sucrose concentrations in tiller bases were significantly higher in  
319 intensively grazed vegetation, supporting the hypothesis of efficient formation of reserves for  
320 plants submitted to intensive grazing (H1). The similar patterns obtained for sucrose and  
321 fructans suggest that both NSCs are involved in the delayed response to grazing intensity. For  
322 the five study grass species, sucrose could thus represent an alternative form of carbohydrate

323 storage. Fulfilling transport function and being easily hydrolysable (Salerno and Curatti,  
324 2003; Amiard *et al.*, 2004), this disaccharide could represent a mobile and “ready to use”  
325 source of carbon, potentially advantageous in a context of intensive grazing. Alternatively,  
326 these higher sucrose concentrations for plants previously submitted to intensive grazing could  
327 reflect a higher rate of photosynthetic assimilation during early spring under intensive  
328 grazing.

329 Several studies have already reported higher NSC concentrations in roots or storage organs of  
330 grass species in response to mowing, either by the end of the growing season (Bartoš *et al.*,  
331 2011) or, as in our study, at the beginning of the following growing season (Baptist *et al.*,  
332 2013; Benot *et al.*, 2013b). However, while NSC concentrations provide a reliable insight in  
333 the investment to storage at the tiller level (i.e., the ratio of biomass allocated to storage), the  
334 effects of grazing (i) on either NSC concentrations or quantities and (ii) at either the tiller or  
335 the tussock level must be considered distinctly (Bartoš *et al.*, 2011; Janeček *et al.*, 2015) as  
336 they may reflect different effects of defoliation on NSC pools. Similar NSC concentrations in  
337 tillers may hide differences in NSC quantities either at the tiller level (linked to differences in  
338 tiller biomass) or at the tussock level (linked to differences in tiller number). Several studies  
339 indeed reported negative effects of experimental defoliation or mowing on tiller biomass or  
340 tiller number at the tussock level (e.g., Donaghy and Fulkerson, 1998; Cullen *et al.*, 2006  
341 Bartoš *et al.*, 2011; Baptist *et al.*, 2013; Benot *et al.*, 2013b). Similarly, we cannot exclude that  
342 the observed inter-specific differences in NSC concentrations can be correlated to inter-  
343 specific differences in the number of tillers per tussock, as suggested by previous  
344 measurements in undisturbed experimental garden conditions for the five study species (see  
345 Fig. S2 in Benot *et al.*, 2013a) or even in tiller turnover rates (as observed in *A. stolonifera*  
346 and *L. perenne* by Bullock *et al.*, 1994). Although we did not record tiller dynamics in the  
347 field, the absence of correlation between the number of tillers per tussock and experimental



348 defoliation severity previously observed for the five study species (Benot *et al.*, 2013a)  
349 suggests that grazing intensity might not have any effect on tiller density for those species.  
350 Similarly, Bullock *et al.* (1994) showed no effect of spring and summer grazing (ie., same  
351 grazing period as in the present study) on tiller densities for *L. perenne* and *A. stolonifera*, but  
352 these results indeed masked summer grazing effects on tiller turnover rates, suggesting  
353 increased tissue renewing under intense grazing. As advocated by Bartoš *et al.* (2011),  
354 considering both concentrations and quantities at the tiller as well as the tussock levels could  
355 provide deeper understanding of grazing effects on carbon economy and plant response to  
356 aboveground biomass loss.

357

#### 358 **Just after a grazing season: a predominant species effect on NSC concentrations**

359 October sampling occurred ten days after the end of the grazing season. At that date, the  
360 composition and concentrations of NSC pools in tiller bases were primarily constrained by the  
361 species rather than by grazing intensity. Such inter-specific differences in carbohydrate  
362 allocation to storage have already been demonstrated for sets of grasses as well as non-grass  
363 species in leaves (Chatterton *et al.*, 1989), roots and diverse stem or root-derived storage  
364 organs (e.g., Janeček and Klimešová 2014). They may be influenced, at least partly, by inter-  
365 specific differences in leaf turnover rates and energy requirements for tissue renewing (Cullen  
366 *et al.*, 2006, Gastal *et al.*, 2010). The absence of difference in NSC concentrations (except  
367 fructose) between moderate and intensive grazing suggested that plants were able to  
368 efficiently restore a functional pool of reserves, regardless of the grazing intensity. Yet,  
369 intensive grazing is likely to generate more severe and/or frequent defoliation than moderate  
370 grazing, which could deplete more completely fructan pools (Beaulieu *et al.*, 1997; Fulkerson  
371 and Donaghy, 2001; Kleijn *et al.*, 2005; Lasseur *et al.*, 2007). In a previous study on two  
372 *L. perenne* varieties, Lasseur *et al.* (2007) observed that frequently defoliated plants showed a

373 higher increase of fructan synthesizing enzyme activities during the period of reserve  
374 replenishment. In line with these findings, our results suggest that plants growing under  
375 intensive grazing may efficiently synthesize NSC reserves between two consecutive  
376 defoliation events (Lee *et al.*, 2010). The higher concentrations of fructose, which is the  
377 product of fructan breakdown by fructan exohydrolase (FEH), under intensive than moderate  
378 grazing ten days after the last defoliation suggested that fructan mobilization efficiency was  
379 enhanced by grazing intensity. Indeed, it is well known that FEH activity is up-regulated after  
380 defoliation (Morvan-Bertrand *et al.*, 1999a, 2001) and may remain at a high level during even  
381 more than ten days following defoliation (Lasseur *et al.*, 2007). Altogether, these results  
382 support the hypothesis of both efficient reserve remobilization and, thereafter, efficient  
383 synthesis resumption for plants submitted to intensive grazing (H2).

384

### 385 **Potential complex grazing effects on NSC concentrations**

386 Grazing is a complex factor not only directly affecting plants through defoliation, but also  
387 through other direct or indirect (e.g., modification of plant environment) effects on plants.  
388 Thus, the NSC patterns recorded in the present study could reflect other grazing effects on  
389 plant physiology or competitive environment. Firstly, grazing intensity might reduce energy  
390 investment in seed production, which is expected to divert resources from vegetative storage  
391 (Chapin *et al.*, 1990; Crone *et al.*, 2009). Although we paid attention to collect exclusively  
392 vegetative tillers from non-flowering clonal fragments, such indirect effects of grazing on  
393 NSC reserve concentrations cannot fully be excluded. Secondly, Liu *et al.* (2012) showed on  
394 the grass *Lymus chinensis*, that sheep saliva addition to clipped plants decreased fructan  
395 concentrations while favouring glucose and fructose accumulation in plant organs during ten  
396 days following treatments. This suggests possible effects of grazing intensity on fructose  
397 concentration through animal saliva. Finally, grazing might also modify the available

398 resources, both in terms of soil nutrients (Rossignol *et al.*, 2011) and light. For instance, by  
399 opening canopy, grazing could limit shading by taller plants and indirectly enhance  
400 photosynthesis and carbohydrate storage in smaller ones. Inter-specific differences in NSC  
401 concentrations in the present study could thus reflect differences in access to light. All of  
402 these potential indirect effects of grazing on reserve making were beyond the scope of the  
403 present study but should deserve more attention in the future.

404

#### 405 **Conclusion and perspectives**

406 The mechanisms of reserve storage regulation and remobilization in response to defoliation  
407 have seemed to be clearly understood for long (see Introduction). Yet, two recent studies  
408 highlighted that carbon storage may in fact be more complex than we do believe (Martínez-  
409 Vilalta *et al.*, 2016; Klimešová *et al.*, 2017). Pathways and organs (e.g., source – sinks  
410 relationships) involved in carbon storage and remobilization in response to disturbance, as  
411 well as their timing and the effect of environmental conditions still remain to be clearly  
412 identified. As expected, our study showed that the concentrations of fructans and sucrose were  
413 significantly higher under intensive than moderate grazing in April (H1), whereas grazing  
414 intensity did not affect NSC concentrations, except for fructose, in October (H2). These  
415 results thus suggest that the composition and the concentrations of NSCs in tiller bases should  
416 be considered to better understand plant responses to grazing. Although field experiments may  
417 suffer from some limitations, for instance due to little spatial or temporal repetitions, we are  
418 convinced that they will help building a stronger understanding of plant carbon economy.

419

420

421

422

## 423 **Authors' contribution**

424 MLB, CM and AB conceived the idea. MLB, CM and JH collected plants in the field. MLB,  
425 AMB, JH, CS, MLD and MPP conceived and performed laboratory analyses. MLB and CM  
426 carried out statistical analyses. All authors discussed the results. MLB, AMB, CM, CS, MPP  
427 and AB wrote the manuscript.

428

## 429 **Funding**

430 This work is a publication from the DIVHERB Project (French national program ECOGER  
431 funded by the *Institut National de la Recherche Agronomique*) and a contribution to GDR  
432 2574 "TRAITS". MLB was funded by a grant from French MESR.

433

## 434 **Acknowledgements**

435 The authors thank N. Renaud and F. Ramel for helpful comments and contribution in the field  
436 and at lab. We are also grateful to the "Parc Naturel Régional" (PNR) du Marais poitevin and  
437 the Town Council of Magnils-Reigniers for supporting ecological research in commonly-  
438 owned grasslands.

439

## 440 **References**

- 441 Albrecht, G., Biemelt, S., Baumgartner S., 1997. Accumulation of fructans following oxygen deficiency stress in  
442 related plant species with different flooding tolerances. *New Phytologist* 136, 137-144.
- 443 Amiard, V., Morvan-Bertrand, A., Cliquet, J.-B., Billard, J.-P., Huault, C., Sandström, J.-P., Prud'homme, M.-P.,  
444 2004. Carbohydrate and amino acid composition in phloem sap of *Lolium perenne* L. before and after  
445 defoliation. *Canadian Journal of Botany* 82, 1594-1601.
- 446 Amiaud, B., Bouzillé, J.-B., Tournade, F., Bonis, A., 1998. Spatial patterns of soil salinities in old embanked  
447 marshlands in western France. *Wetlands* 18, 482-494.

448 Asaeda, T., Rajapakse, L., Manatunge, J., Sahara, N., 2006. The effect of summer harvesting of *Phragmites*  
449 *australis* on growth characteristics and rhizome resource storage. *Hydrobiologia* 553, 327-335.

450 Bakker, C., Blair, J.M., Knapp, A.K., 2003. Does resource availability, resource heterogeneity or species  
451 turnover mediate changes in plant species richness in grazed grasslands? *Oecologia* 137, 385-391.

452 Baptist, F., Secher-Fromell, H., Viart-Crétat, F., Aranjuelo, I., Clément, J.-C., Crème, A., Desclos, M., Laine, P.,  
453 Nogues, S., Lavorel, S., 2013. Carbohydrate and nitrogen stores in *Festuca paniculata* under mowing  
454 explain dominance in subalpine grasslands. *Plant Biology* 15, 395-404.

455 Bartoš, M., Janeček, Š., Klimešová, J., 2011. Effect of mowing and fertilization on biomass and carbohydrate  
456 reserves of *Molinia caerulea* at two organizational levels. *Acta Oecologica* 37, 299-306.

457 Beaulieu, J., Gauthier, G., Rochefort, L., 1997. The growth response of graminoid plants to goose grazing in a  
458 High Arctic environment. *Journal of Ecology* 84, 905-914.

459 Benot, M.-L., Mony, C., Lepš, J., Penet, L., Bonis, A., 2013a. Are clonal traits and their response to defoliation  
460 good predictors of grazing resistance? *Botany* 91, 62-68.

461 Benot, M.-L., Saccone, P., Vicente, R., Pautrat, E., Morvan-Bertrand, A., Decau, M.-L., Grigulis, K.,  
462 Prud'homme, M.-P., Lavorel, S., 2013b. How extreme summer weather may limit control of *Festuca*  
463 *paniculata* by mowing in subalpine grasslands. *Plant Ecology & Diversity* 6, 393-404.

464 Bloom, A.J., Chapin, F.S. III, Mooney, H.A., 1985. Resource limitation in plants – An economic analogy. *Annual*  
465 *Review of Ecology and Systematics* 16, 363-392.

466 Bonis, A., Bouzillé, J.-B., Amiaud, B., Loucougaray, G., 2005. Plant community patterns in old embanked  
467 grasslands and the survival of halophytic flora. *Flora* 200, 74-87.

468 Briske, D.D., 1996. Strategies of plant survival in grazed systems: a functional interpretation. In: Hodgson J. and  
469 Illius A.W., (eds) *The Ecology and management of grazing systems*. Wallingford: CAB International, pp.  
470 33-67.

471 Brocklebank, K.J., Hendry, J.A.F., 1989. Characteristics of plant species which store different types of reserve  
472 carbohydrates. *New Phytologist* 112, 255-263.

473 Bullock, J.M., Clear Hill, B., Silvertown J., 1994. Tiller dynamics of two grasses – responses to grazing, density  
474 and weather. *Journal of Ecology* 82, 331-340.

475 Bullock, J.M., Franklin, J., Stevenson, M.J., Silvertown, J., Coulson, S.J., Gregory, S.J., Softs, R., 2001. A plant  
476 trait analysis of responses to grazing in a long-term experiment. *Journal of Applied Ecology* 38, 253-267.

477 Cairns, A.J., Cookson, A., Thomas, B.J., Turner, L.B., 2002. Starch metabolism in the fructan-grasses: patterns of  
478 starch accumulation in excised leaves of *Lolium temulentum* L. *Journal of Plant Physiology* 159, 293-  
479 305.

480 Chapin, F.S. III, Schulze, E.D., Mooney, H.A., 1990. The ecology and economics of storage in plants. *Annual*  
481 *Review of Ecology and Systematics* 21, 423-447.

482 Chatterton, N.J., Harrisson, P.A., Bennett, J.H., Asay, K.H., 1989. Carbohydrate partitioning in 185 accessions of  
483 Gramineae grown under warm and cool temperatures. *Journal of Plant Physiology* 134, 169-179.

484 Crone, E.E., Miller, E., Sala, A., 2009. How do plants know when other plants are flowering? Resource  
485 depletion, pollen limitation and mast-seeding in a perennial wildflower. *Ecology Letters* 12, 1119-1126.

486 Cullen, B.R., Chapman, D.F., Quigley, P.E., 2006. Comparative defoliation tolerance of temperate perennial  
487 grasses. *Grass and Forage Science* 61, 405-412.

488 de Bello, F., Lepš, J., Sebastià, M.T., 2005. Predictive value of plant traits to grazing along a climatic gradient in  
489 the Mediterranean. *Journal of Applied Ecology* 42, 824-833.

490 de Jong, T., van der Meijden, E., 2000. On the correlation between allocation to defence and regrowth in plants.  
491 *Oikos* 88, 503-508.

492 de Mendiburu, F., 2014. agricolae: Statistical Procedures for Agricultural Research. R package version 1.2-1.  
493 <http://CRAN.R-project.org/package=agricolae>

494 de Visser, R., Vianden, H., Schnyder, H., 1997. Kinetics and relative significance of remobilized and current C  
495 and N incorporation in leaf and root growth zones of *Lolium perenne* after defoliation: assessment by <sup>13</sup>C  
496 and <sup>15</sup>N steady-state labelling. *Plant, Cell and Environment* 20, 37-46.

497 Díaz, S., Noy-Meir, I., Cabido, M., 2001. Can grazing response of herbaceous plants be predicted from simple  
498 vegetative traits? *Journal of Applied Ecology* 38: 497-508.

499 Díaz, S., Lavorel, S., McIntyre, S., Falczuk, V., Casanoves, F., Milchunas, D.G., Skarpe, C., Rusch, G.,  
500 Sternberg, M., Noy-Meir, I., Landsberg, J., Zhang, W., Clark, H., Campbell, B.D., 2007. Plant trait  
501 response to grazing – a global synthesis. *Global Change Biology* 13, 313-341.

502 Donaghy, D.J., Fulkerson, W.J., 1998. Priority for allocation of water-soluble carbohydrate reserves during  
503 regrowth of *Lolium perenne* (L). *Grass and Forage Science* 53, 211–218.

504 Fulkerson, W.J., Donaghy, D.J., 2001. Plant-soluble carbohydrate reserves and senescence - key criteria for  
505 developing an effective grazing management system for ryegrass-based pastures: a review. *Australian*  
506 *Journal of Experimental Agriculture* 41, 261–275.

507 Gastal, F., Dawson, A., Thorton, B., 2010. Responses of plant traits of four grasses from contrasting habitats to  
508 defoliation and N supply. *Nutrient Cycling in Agroecosystems* 88, 245-258.

509 Hawker, J.S., 1985. Sucrose. In: Dey P.M. and Dixon R.A. (eds) *Biochemistry of storage carbohydrates in green*  
510 *plants*. London: Academic Press, pp. 1-52.

511 Haydon, M.J., Bell, L.J., Webb, A.A.R., 2011. Interactions between plant circadian clocks and solute transport.  
512 *Journal of Experimental Botany* 62, 2333-2348.

513 Högling, M., Hanslin, H.M., Mortensen, L.M., 2011. Photosynthesis of *Lolium perenne* L. at low temperatures  
514 under low irradiances. *Environmental and Experimental Botany* 70, 297-304.

515 Janeček, Š., Lanta, V., Klimešová, J., Doležal, J., 2011. Effect of abandonment and plant classification on  
516 carbohydrate reserves of meadow plants. *Plant Biology* 13, 243-251.

517 Janeček, Š., Klimešová, J., 2014. Carbohydrate storage in meadow plants and its depletion after disturbance: do  
518 roots and stem-derived organs differ in their roles? *Oecologia* 175, 51-61.

519 Janeček, Š., Bartušková, A., Bartoš, M., Altman, J., de Bello, F., Doležal, J., Latzel, V., Lanta, V., Lepš, J.,  
520 Klimešová, J., 2015. Effects of disturbance regime on carbohydrate reserves in meadow plants. *AoB*  
521 *Plants* 7, plv123.

522 Karban, R., Baldwin, I.T., 1997. *Induced responses to herbivory*. Chicago: University of Chicago Press.

523 Kattge, J., Díaz S., Lavorel, S., Prentice, I.C., Leadley, P., Bönsch, G., Garnier, E., Westoby ,M., Reich, P.B.,  
524 Wright, I.J., Cornelissen, J.H., Violle, C., Harrison, S.P., Van Bodegom, P.M., Reichstein, M., Enquist,  
525 B.J., Soudzilovskaia, N.A., Ackerly, D.D., Anand, M., Atkin, O., Bahn, M., Baker, T.R., Baldocchi, D.,  
526 Bekker, R., Blanco, C.C., Blonder, B., Bond, W.J., Bradstock, R., Bunker, D.E., Casanoves, F.,  
527 Cavender ~~Blare~~ Chambers, J.Q., Chapin, III F.S., Chave, J., Coomes, D., Cornwel, W.K., Craine,  
528 J.M., Dobrin, B.H., Duarte, L., Durka, W., Elser, J., Esser, G., Estiarte, M., Fagan, W.F., Fang, J.,  
529 Fernández ~~Blare~~ Fernández, F., Friedel, A.G., Fine, K.S., Flores ,O., Forner,  
530 N.M., Gallagher, R.V., Green, W.A., Gutierrez, A.G., Hickler, T., Higgins, S.I., Hodgson, J.G., Jalili, A.,  
531 Jansen, S., Joly, C.A., Kerkhoff, A.J., Kirkup, D., Kitajima, K., Kleyer, M., Klotz, S., Knops, J.M.,  
532 Kramer, K., Kühn, I., Kurokawa, H., Laughlin, D., Lee, T.D., Leishman, M., Lens, F., Lenz, T., Lewis,  
533 S.L., Lloyd, J., Llusià, J., Louault, F., Ma, S., Mahecha, MD., Manning, P., Massad, T., Medlyn, B.E.,  
534 Messier, J., Moles, A.T., Müller, SC., Nadrowski, K., Naeem, S., Niinemets, Ü., Nöllert, S., Nüske, A.,  
535 Ogaya, R., Oleksyn, J., Onipchenko, V.G., Onoda, Y., Ordoñez, J., Overbeck, G., Ozinga, W.A., Patiño,  
536 S., Paula, S., Pausas, J.G., Peñuelas, J., Phillips, O.L., Pillar, V., Poorter ,H., Poorter, L., Poschlod, P.,

- 537 Prinzing, A., Proulx, R., Rammig, A., Reinsch, S., Reu, B., Sack, L., Salgado -Negret, B., Saro  
538 Shiodera, S., Shipley, B., Siefert, A., Sosinski, E., Soussana, J., Swaine, E., Swenson, N., Thompson,  
539 K., Thornton, P., Waldram, M., Weiher, E., White, M., White, S., Wright, S.J., Yguel, B., Zaehle, S.,  
540 Zanne AE., Wirth C. 2011. TRY – a global database of plant traits. *Global Change Biology* 17, 2905-  
541 2935.
- 542 Kavanová, M., Gloser, V., 2005. The use of internal nitrogen stores in the rhizomatous grass *Calamagrostis*  
543 *epigejos* during regrowth after defoliation. *Annals of Botany* 95, 457-463.
- 544 Kleijn, D., Treier, U.A., Müller-Schärer, H., 2005. The importance of nitrogen and carbohydrate storage for plant  
545 growth of the alpine herb *Veratrum album*. *New Phytologist* 166, 565-575.
- 546 Klimeš, L., Klimešová, J., 2002. The effects of mowing and fertilization on carbohydrate reserves and regrowth  
547 of grasses: do they promote plant coexistence in species-rich meadows? *Evolutionary Ecology* 15, 363-  
548 382.
- 549 Klimešová, J., Klimeš, L., 2003. Resprouting of herbs in disturbed habitats: is it adequately described by  
550 Bellingham - Sparrow's model? *Oikos* 103, 225-229.
- 551 Klimešová, J., Janeček, S., Bartušková, A., Bartoš, M., Altman, J., Doležal, J., Lanta, V., Latzel, V., 2017. Is the  
552 scaling relationship between carbohydrate storage and leaf biomass in meadow plants affected by the  
553 disturbance regime? *Annals of Botany* 120, 979-985.
- 554 Kobe, R.K., 1997. Carbohydrate allocation to storage as a basis of interspecific variation in sapling survivorship  
555 and growth. *Oikos* 80, 226-233.
- 556 Kohler, F., Gillet, F., Gobat, J.M., Buttler, A., 2004. Seasonal vegetation changes in mountain pastures due to  
557 simulated effects of cattle grazing. *Journal of Vegetation Science* 15, 143-150.
- 558 Kühn, I., Durka, W., Klotz, S., 2004. BioFlor - a new plant-trait database as a tool for plant invasion ecology.  
559 *Diversity and Distribution* 10, 363-365.
- 560 Lasseur, B., Lothier, J., Morvan-Bertrand, A., Escobar-Gutiérrez, A., Humphreys, M.O., Prud'homme, M.-P.,  
561 2007. Impact of defoliation frequency on regrowth and carbohydrate metabolism in contrasting varieties  
562 of *Lolium perenne*. *Functional Plant Biology* 34, 418-430.
- 563 Lee, J.M., Sathish, P., Donaghy, D.J., Roche, J.R., 2010. Plants modify biological processes to ensure survival  
564 following carbon depletion: a *Lolium perenne* model. *PLoS One* 5, e12306.
- 565 Liu, J., Wang, L., Wang, D., Bonser, S.P., Sun, F., Zhou, Y., Gao, Y., Teng, X., 2012. Plants can benefit from  
566 herbivory: stimulatory effects of sheep saliva on growth of *Leymus chinensis*. *PLoS One* 7, e29259.



567 Loucougaray, G., Bonis, A., Bouzillé, J.-B., 2004. Effects of grazing by horses and/or cattle on the diversity of  
568 coastal grasslands in western France. *Biological Conservation* 116, 59-71.

569 Manner, D.J., 1985. Starch. In: Dey P.M. and Dixon R.A. (eds) *Biochemistry of storage carbohydrates in green*  
570 *plants*. London: Academic Press, pp. 149-203.

571 Marion, B., Bonis, A., Bouzillé, J.-B., 2010. How much grazing-induced heterogeneity impact plant diversity  
572 and richness in wet grasslands? *Ecoscience* 17, 229-239.

573 Martínez-Vilalta, J., Sala A., Asensio D., Galiano, L., Hoch, G., Palacio, S., Piper, F., Lloret, F., 2016. Dynamics  
574 of non-structural carbohydrates in terrestrial plants: a global synthesis. *Ecological Monographs* 86: 495–  
575 516.

576 Maschinski, J., Whitham, T.G., 1989. The continuum of plant responses to herbivory: the influence of plant  
577 association, nutrient availability, and timing. *American Naturalist* 134, 1-19.

578 McNaughton, S.J., 1983. Compensatory plant growth as a response to herbivory. *Oikos*, 40, 329-336.

579 Ménard, C., Ducan, P., Fleurance, G., George, J.Y., Lila, M., 2002. Comparative foraging and nutrition in horses  
580 and cattle in Europe wetlands. *Journal of Applied Ecology* 38, 120-133.

581 Morvan-Bertrand, A., Boucaud, J., Prud'homme, M.-P., 1999a. Influence of initial levels of carbohydrates,  
582 fructans, nitrogen, and soluble proteins on regrowth of *Lolium perenne* L. cv. Bravo following defoliation.  
583 *Journal of Experimental Botany* 50, 1817-1825.

584 Morvan-Bertrand, A., Pavis, N., Boucaud, J., Prud'homme, M.-P., 1999b. Partitioning of reserve and newly  
585 assimilated carbon in roots and leaf tissues of *Lolium perenne* during regrowth after defoliation:  
586 assessment by <sup>13</sup>C steady-state labelling and carbohydrate analysis. *Plant, Cell and Environment* 22,  
587 1097-1108.

588 Morvan-Bertrand, A., Boucaud, J., Le Saos, J., Prud'homme, M.-P., 2001. Roles of fructans from leaf sheaths  
589 and from the elongating leaf bases in the regrowth following defoliation of *Lolium perenne* L. *Planta* 213,  
590 109-120.

591 Pollock, C.J., and Cairns, A.J., 1991. Fructan metabolism in grasses and cereals. *Annual Review of Plant*  
592 *Physiology and Plant Molecular Biology* 42, 77-101.

593 R Development Core Team, 2014. R: A language and environment for statistical computing. Vienna, Austria: R  
594 Foundation for Statistical Computing. URL <http://www.R-project.org/>

595 Richards, J.H., 1993. Physiology of plants recovering from defoliation. In: Baker M.J. (ed) *Proceedings of the*  
596 *XVII International Grassland Congress*. Wellington: SIR Publishing, pp. 85-94.

597 Rosa, M., Prado, C., Podazza, G., Interdonato, R., González, J.A., Hilal, M., Prado, F.E., 2009. Soluble sugars—  
598 Metabolism, sensing and abiotic stress. A complex network in the life of plants. *Plant Signaling &*  
599 *Behavior* 4, 388-393.

600 Rossignol, N., Bonis, A., Bouzillé, J.-B., 2006. Relationships between grazing pattern and vegetation structure  
601 on the spatial variations of net N mineralisation in a wet grassland. *Applied Soil Ecology* 31, 62-72.

602 Rossignol, N., Bouzillé, J.-B., Bonis, A., 2011. Grazing-induced vegetation patchiness controls net N  
603 mineralization rate in a semi-natural grassland. *Acta Oecologica* 37, 290-297.

604 Sala, O.E., Oesterheld, M., León, R.J.C., Soriano, A., 1986. Grazing effects upon plant community structure in  
605 subhumid grasslands of Argentina. *Vegetatio* 67, 27-32.

606 Salerno, G.L., Curatti, L., 2003. Origin of sucrose metabolism in higher plants: when, how and why? *Trends in*  
607 *Plant Science* 8, 63–69.

608 Schnyder, H., de Visser, R., 1999. Fluxes of reserve-derived and currently assimilated carbon and nitrogen in  
609 perennial ryegrass recovering from defoliation. The regrowing tiller and its component functionally  
610 distinct zones. *Plant Physiology* 119, 1423-1435.

611 Shewmaker, G.E., Mayland, H.F., Roberts, C.A., Harrison, P.A., Chatterton, N.J., Sleper, D.A., 2006. Daily  
612 carbohydrate accumulation in eight tall fescue cultivars. *Grass and Forage Science* 61, 413-421.

613 Smith, A.M., Stitt, M., 2007. Coordination of carbon supply and plant growth. *Plant, Cell and Environment* 30,  
614 1126-1149.

615 Steen, E., Larsson, K., 1986. Carbohydrates in roots and rhizomes of perennial grasses. *New Phytologist* 104,  
616 339-346.

617 Stowe, K.A., Marquis, R.J., Hochwender, C.G., Simms, E.L., 2000. The evolutionary ecology of tolerance to  
618 consumer damage. *Annual Review of Ecology and Systematics* 31, 565-595.

619 Sulmon, C., Gouesbet, G., Ramel, F., Cabello-Hurtado, F., Penno, C., Bechtold, N., Couée, I., El Amrani, A.,  
620 2011. Carbon Dynamics, Development and Stress Responses in *Arabidopsis*: Involvement of the APL4  
621 Subunit of ADPGlucose Pyrophosphorylase (Starch Synthesis). *PlosOne* 16, e26855 .

622 Suzuki, J.I., Hutchings, M.J., 1997. Interactions between shoots in clonal plants and the effect of stored resources  
623 on the structure of shoot populations. In: de Kroon H. and van Groenendael J. (eds) *The ecology and*  
624 *evolution of clonal plants*. Leiden: Backhuys Publishers, pp. 311-329.

625 van der Meijden, E., Wijn, M., Verkaar, H., 1988. Defence and regrowth, alternative plant strategies in the  
626 struggle against herbivores. *Oikos* 51, 355-363.

- 627 van der Meijden, E., de Boer, N.J., van der Veen-van Wijk, C.A.M., 2000. Pattern of storage and regrowth in  
628 ragwort. *Evolutionary Ecology* 14, 439-455.
- 629 Veen, G.F., Blair, J.M., Smith, M.D., Collins, S.L., 2008. Influence of grazing and fire frequency on small-scale  
630 plant community structure and resource variability in native tallgrass prairie. *Oikos* 117, 859-866.

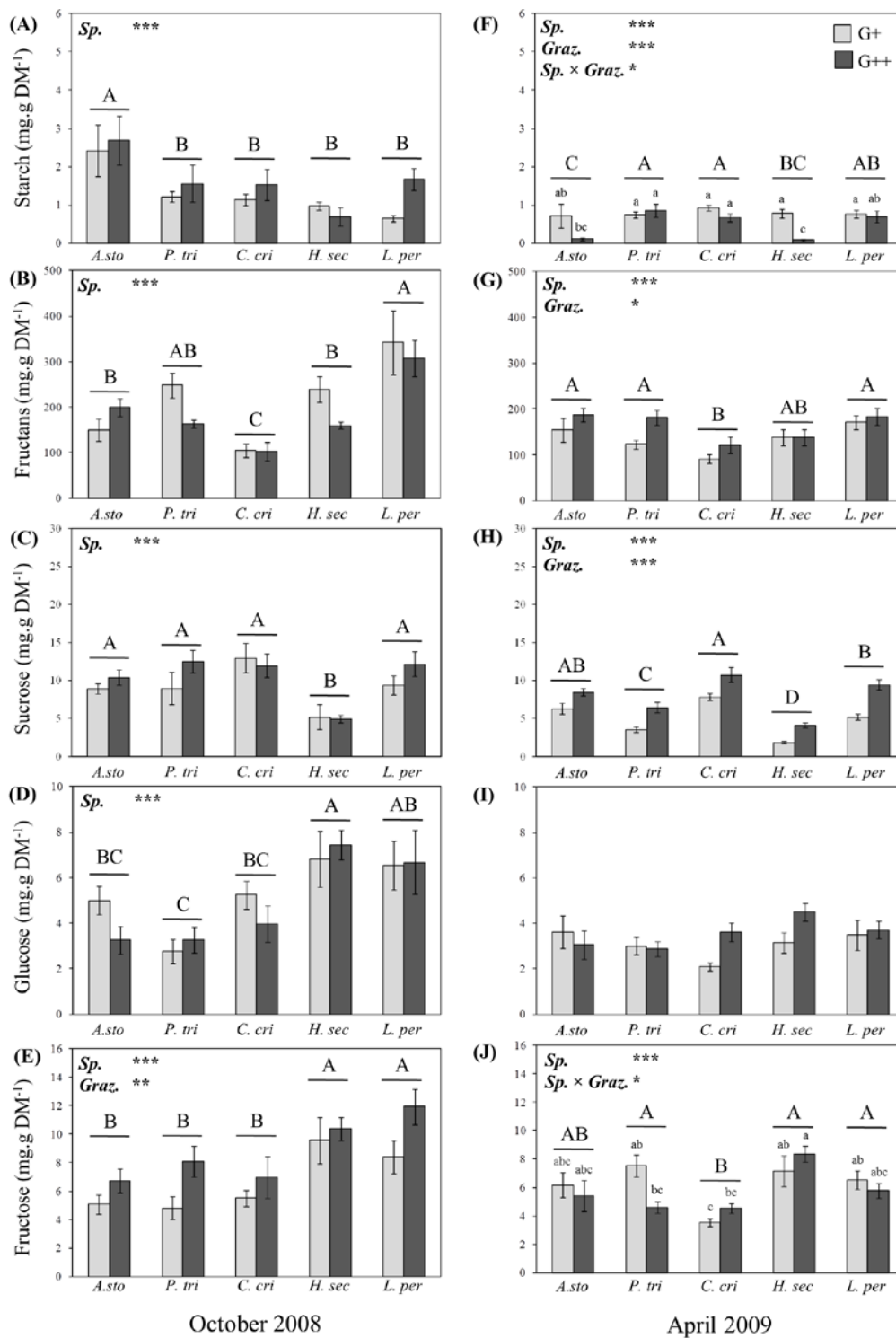
**Table 1** Mean ( $\pm$  SE) species relative cover (%) under moderate (G+) and intensive (G++) grazing (n=10 sampling plots per grazing intensity) and grazing tolerance retrieved TRY database (Kattge *et al.* 2011; original source: BiolFlor database, Kühn *et al.* 2004).

	<i>A. stolonifera</i>	<i>P. trivialis</i>	<i>C. cristatus</i>	<i>H. secalinum</i>	<i>L. perenne</i>
Cover in G+ (%)	8.7 $\pm$ 3.5	10.5 $\pm$ 2.2	16.2 $\pm$ 2.2	7.49 $\pm$ 2.2	0.4 $\pm$ 0.3
Cover in G++ (%)	5.0 $\pm$ 2.7	9.4 $\pm$ 2.6	17.0 $\pm$ 3.51	12.3 $\pm$ 2.8	17.7 $\pm$ 1.6
Grazing tolerance*	9	6	7	7	8

\* ranging from 1 (intolerant) to 9 (very tolerant to grazing).

**Table 2** Result of the ANOVA testing for the impact of sampling date (*Date*), species (*Sp.*) and grazing intensity (*Graz.*) and their interactions on non-structural carbohydrate (NSC) concentrations in tiller bases. Significant P-values (<0.05) are in bold. Data transformations for the models are indicated (log: log-transformed data, asin-sqrt: arcsine square root-transformed data).

	Starch		Fructans		Sucrose		Glucose		Fructose		
	df	F	P	F	P	F	P	F	P	F	P
		<i>n</i> = 158		<i>n</i> = 153		<i>n</i> = 153		<i>n</i> = 153		<i>n</i> = 153	
Date	1	53.81	<b>&lt;0.001</b>	14.72	<b>&lt;0.001</b>	71.25	<b>&lt;0.001</b>	31.21	<b>&lt;0.001</b>	19.99	<b>&lt;0.001</b>
Sp.	4	4.85	<b>0.001</b>	19.59	<b>&lt;0.001</b>	48.16	<b>&lt;0.001</b>	8.49	<b>&lt;0.001</b>	12.66	<b>&lt;0.001</b>
Graz.	1	1.98	0.16	0.91	0.34	43.42	<b>&lt;0.001</b>	0.10	0.76	3.58	0.061
Date × Sp.	4	7.98	<b>&lt;0.001</b>	2.95	<b>0.023</b>	3.07	<b>0.019</b>	3.28	<b>0.013</b>	2.33	0.059
Date × Graz.	1	9.74	<b>0.002</b>	5.21	<b>0.024</b>	11.37	<b>&lt;0.001</b>	3.29	0.072	9.49	<b>0.002</b>
Sp. × Graz.	4	3.76	<b>0.006</b>	1.53	0.20	1.88	0.12	1.51	0.20	0.40	0.81
Date × Sp. × Graz.	4	0.75	0.56	1.48	0.21	0.76	0.55	1.19	0.32	2.69	<b>0.034</b>
Transformation		asin sqrt		log		log		asin sqrt		asin sqrt	



**Figure 1** Mean ( $\pm$  SE) non-structural carbohydrate (NSC) concentrations (mg g DM<sup>-1</sup>) in tiller bases in October 2008 (A-E) and April 2009 (F-J). Species are ranged according to increasing difference in relative cover between G+ and G++ (see Table 1). Only significant effects of species (*Sp.*), grazing

intensity (*Graz.*) and their interaction are indicated: \*  $P < 0.05$ , \*\*  $P < 0.01$ , \*\*\*  $P < 0.001$ . Different letters indicate significant differences. G+ moderate grazing, G++ intensive grazing.