

Development of On-Line Solid-Phase Extraction-Liquid Chromatography Coupled with Tandem Mass Spectrometry Method to Quantify Pharmaceutical, Glucuronide Conjugates and Metabolites in Water

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1	Development of on-line solid-phase extraction-liquid chromatography								
2	coupled with tandem mass spectrometry method to quantify pharmaceutical,								
3	glucuronide conjugates, and metabolites in water.								
4									
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14 Abstract:

15 The present work describes the development of an analytical method, based on automated on-line solid phase extraction followed by ultra-high-performance liquid chromatography 16 coupled with tandem mass spectrometry (SPE-LC-MS/MS) for the quantification of 37 17 pharmaceutical residues, covering various therapeutic classes, and some of their main 18 metabolites, in surface and drinking water. A special attention was given to some 19 glucuronide conjuguates and metabolites of active subtances. Multiple Reaction Monitoring 20 21 (MRM) was chosen and two transitions per compound are monitored (quantification and 22 confirmation transitions). Quantification is performed by standard addition approach to correct matrix effect. The method provides limit of quantification inferior to 20 ng.L⁻¹ for all 23 compounds. The methodology was successfully applied to the analysis of surface water and 24 drinking water of 8 drinking water treatment plant in west of France. The highest drug 25 concentrations in surface water and drinking water were reported for ketoprofen, 26 27 hydroxyibuprofen, acetaminophen, caffeine and danofloxacin.

28

Key words: pharmaceuticals, automated on-line solid phase extraction, liquid
 chromatography, tandem mass spectrometry, water analysis

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31 **1. Introduction**

Pharmaceuticals are an important group of emerging contaminants in the environment [1]. 32 In recent years many reports have been made on the occurrence of the large, differentiated 33 group of pharmaceuticals in wastewater, surface water, ground water and drinking water in 34 many countries [2-9]. After administration, most pharmaceuticals are not completely 35 36 metabolized. The unmetabolized parent pharmaceutical and some metabolites are 37 subsequently excreted from the body via urine and faeces [10]. Reports have shown that many pharmaceuticals do not totally degrade during conventional wastewater treatment 38 [11,12]. The concentrations of individual compounds in wastewater, surface water, ground 39 water and drinking water are typically in the range of ng/L to μ g/L. The effect on long-term 40 pharmaceutical residues in aquatic environments remains largely unknown. In addition, the 41 42 risks to the environment are evaluated for a particular drug, while we find a mixture of all these compounds in aquatic environments. Studies have shown that combinations of drugs 43 may be more powerful than the simple addition of two drugs individually toxic effects [13-44 14]. 45

Wastewater effluent is a major source for the input of pharmaceuticals to the environment [11;12], which can then migrate through water systems and into source water intended for drinking water supplies. Advanced wastewater treatment processes have been shown to significantly reduce the concentrations of emerging contaminants. However, some compounds are not completely removed even if treatment techniques are used [15]. Moreover, most of the WWTP do not include these specifically designed treatment units.

In this context, sensitive analytical methods allowing the quantification of many pollutants at trace concentration is essential. Solid Phase Extraction (SPE) is the most commonly used technique to prepare sample before analysis. SPE allows the concomitance of analyte 55 concentration and interferences removal [16;17]. To date, most of the published multiresidue methods for the determination of ultra traces of pharmaceuticals compounds in 56 surface and drinking water use off-line SPE followed by gas chromatography mass 57 spectrometry (GC-MS) or by liquid chromatography-tandem mass spectrometry (LC-58 MS/MS) [2-5;7;9;12]. However, On-line Solid Phase Extraction is an emerging method for 59 60 analysis of the trace compounds of organic micropollutants (reactive drugs, pesticides...). 61 This technique has many advantages: saving time, automated method, reproducibility, very low solvent consumption, small sample handling, SPE cartridges reuse... [17]. The cartridges 62 used to concentrate pharmaceuticals residues are usually Oasis[™] HLB or hydrophobic resins. 63 [18;19]. This technique is generally coupled to liquid chromatography with UV, MS or MS/MS 64 detector with reversed phase column [20-24]. 65

The objectives of this work has been to develop a fully automated method to analyze a 66 number of target compounds belonging to different therapeutical classes and some by 67 product using on-line SPE directly coupled to liquid chromatography tandem mass 68 spectrometry (LC–MS/MS). This analytical technique limits matrix effect impact. However 69 remaining, interfering species can affect the analytical train, especially natural organic 70 71 matter may coeluate with targeted compounds which leads to a signal disturbance causing 72 over/underestimation or false positive results, or some compounds may react with targeted 73 molecules during sampling and storage [25].

This method was evaluated in different water matrices: UltraPure Water (UPW) to develop
the analytical method, surface water and drinking water for validation.

76

- 77 2. Material and methods78
- 79 *2.1. Compound selection*

81 32 pharmaceuticals and 3 metabolites and 2 glucuronide conjugates were selected for this study (Table 1 and Table S1). These molecules were chosen based on the following criteria: i) 82 selected compounds should exhibit a variety of physical properties, such as functional 83 groups and polarity, ii) they should represent of a diversity of pharmaceutical classes, iii) 84 high frequencies of environmental occurrence, iv) low removal efficiencies by drinking water 85 86 and wastewater treatment techniques in France or others countries [2-9]. Table 1 lists the 37 87 molecules selected for our study and their optimized parameters for quantification, chemical structure is provided in the figure S1 in Supporting Information. Thereafter, the molecules 88 will be called by the short identifiers which are given in the table 1. The pharmaceutical 89 classes represented are cardiovascular drugs, anticancer agents, human or veterinary 90 antibiotics, neuroleptics, non-steroidal anti-inflammatory drugs and hormones. 91

92

93 2.2. Pharmaceutical standards and reagents

All pharmaceutical compounds have minimum 90% purity, used as received in solid form and were obtained from Sigma Aldrich (FRANCE). Ultra pure water (UPW) was delivered by a ElgaPureLab System (resistivity 18.2 M Ω .cm, COT< 50µg C/L). Chromatographic and SPE solvents, acetonitrile (ACN) with or without 0.1 % formic acid (FA) and methanol (MeOH) were purchased from JT Baker (LC-MS grade) and were used in association with UPW in also or not with 0.1 % formic acid.

All concentrated stock solution of individual pharmaceuticals were prepared in methanol with a concentration of 500 mg.L⁻¹ and stored at $-20 \circ$ C. The mixed spiking solutions were prepared in methanol at 500 µg.L⁻¹ and stored at 4°C during 15 days maximum. This mixed spiking solution is daily diluted in water to obtained 500 ng.L⁻¹ before use for standard

104	addition.	Concentrations	prepared	for	analytical	development	and	to	quantify	the	target
105	compoun	ds in the differer	nt matrices	s are	e: 5, 10, 20	, 50, 100, 250 a	and 5	500	ng.L ⁻¹ .		

107 2.3. On-line solid phase extraction and liquid chromatography

108 The analytical system consists of an automated SPE sampler coupled with an LC-MS/MS. The 109 online extraction was carried out using a 2777 autosampler equipped with two parallel Oasis[™] HLB cartridge (Direct Connect HP 20µm, 2.1x30mm) working sequentially. The 110 switching from the loading flow pattern, to elution, then conditioning and back to loading is 111 performed using two six positions $Everflow^{TM}$ values. Loading eluent (UPW) and conditioning 112 eluent (methanol) were provided by a quaternary pump (Acquity[™] QSM). Elution of the 113 114 analytes from the SPE cartridge to LC system was achieved by connected the cartridge to the 115 inlet of the separation column and using the initial chromatographic elution solution.

Separation was carried out using a reversed phase column (AcquityTM BEH C18, 100 mm x 2.1 mm ID, 17 μ m) placed in an oven (45°C). The elution gradient was produced by a binary pump (AcquityTM BSM) and was optimized and will be described later in the manuscript.

119

120 *2.4. Mass spectrometry*

The mass spectrometer (Quattro Premier, MicromassTM) operates with the following conditions: cone gas (N2, 50 L.h⁻¹, 120 °C), desolvation gas (N2, 750 L.h⁻¹, 350 °C), collision gas (Ar, 0.1 mL.min⁻¹), capillary voltage (3000V). The ionization source of the mass spectrometer is an electrospray (ESI) used either in the positive or the negative mode according to pharmaceutical compounds structure (table 1). All the analysis, are made in "multiple reaction monitoring" (MRM) mode, the parent ion from the ESI source is selected in the first quadrupole (pseudomolecular ion in most cases) and fragmented in the collision

- 128 cell. One or more fragments (quantification ion and, when available, confirmation ions) are
- 129 then selected by the third quadrupole before being detected by a photomultiplier. This
- 130 mode allows high sensitivity and selectivity.

Table 1: List of the 35 pharmaceuticals with pharmaceutical class Molecule (short identifier), N°CAS, MW (g/mol), formula, mass parameter and retention time

Pharmaceutical class	Molecule (short identifier)	N°CAS	MW (g/mol)	formula of the active substance	ESI	Parents ion	Daughter ion(Q)	Cones (V)	Collisions (V)	Confirmation ion	Collisions (V)	Dwell time (ms)	Tr (min)
	Amlodipin (AML)	111470-99-6	567.05	C ₂₀ H ₂₅ ClN ₂ O ₅	+	409.6	238.1	18	11	409.6	13	50	4.03
	Atenolol (ATE)	29122-68-7	266.34	C ₁₄ H ₂₂ N ₂ O ₃	+	267	145	34	26	74	23	50	1.18
	Losartan (LOS)	124750-99-8	461	C ₂₂ H ₂₃ CIN ₆ O	+	423.6	405.2	30	12	207	22	50	4.25
	Naftidrofuryl (NAF)	03200-6-4	473.56	C ₂₄ H ₃₃ NO ₃	+	384.6	99.7	40	21	84.7	25	50	4.29
Cardiovascular drugs	Pravastatin (PRA)	81131-70-6	446.51	C ₂₃ H ₃₆ O ₇	-	423.2	100.6	34	23	321.1	16	50	2.63
	Propanolol (PRO)	525-66-6	259.4	C ₁₆ H ₂₁ NO ₂	+	260.2	116	34	18	183	18	50	3.33
	Gemfibrozil (GEM)	25812-30-0	250.33	C ₁₅ H ₂₂ O ₃	-	249	121	34	23			50	4.95
	Trimetazidin (TRI)	13171-25-0	339.26	$C_{14}H_{24}CI_2N_2O_3$	+	267.4	180.9	21	16	165.8	26	50	1.18
	Tamoxifen (TAM)	10540-29-1	371.5	C ₂₆ H ₂₉ NO	+	372.5	72	45	14			50	5.42
anticancer agent	Hydroxytamoxifen (OH-TAM)	68047-06-3	387.2	C ₂₆ H ₂₉ NO ₂	+	388.2	72	45	14			50	4.58
	Ifosfamide (IFO)	3778-73-2	261	C ₇ H ₁₅ Cl ₂ N ₂ O ₂ P	+	261.02	153.95	25	22	92.04	25	75	3
	Doxycycline (DOX)	24390-14-5	512.94	$C_{22}H_{24}N_2O_8$	+	445.5	428.2	30	18	153.8	28	50	2.95
	Erythromyicin (ERY)	114-07-8	769.96	C ₃₇ H ₆₇ NO ₁₃	+	734.2	158	28	30	576.2	19	50	3.68
Human Antibiotic	Ofloxacin (OFX)	82419-36-1	361.37	C ₁₈ H ₂₀ FN ₃ O ₄	+	362	318	34	19	261	28	80	1.35
	Sulfaméthoxazole (SUL)	723-46-6	253.278	C ₁₀ H ₁₁ N ₃ O ₃ S	+	254	92	26	28	156	16	50	2.74
	Trimetoprime (TRP)	738-70-5	290.3	$C_{14}H_{18}N_4O_3$	+	291.2	230	24	24	261.1	26	50	1.18
	Danofloxacin (DANO)	112398-08-0	357.38	C ₁₉ H ₂₀ FN ₃ O ₃	+	358.5	314	35	19	283	25	50	1.53
Vetering in Autibiatia	Lincomycin (LINCO)	859-18-7	461.37	C ₁₈ H ₃₄ N ₂ O ₆ S	+	407.6	125.9	40	28	359.3	18	50	1.23
vetermanan Antibiotic	Sulfadimerazine (SFZ)	57-68-1	278.33	$C_{11}H_{12}N_4O_2S$	+	279.4	185.9	29	16	91.7	26	50	1.91
	Tylosin (TYL)	74610-55-2	1066.19	C ₄₆ H ₇₇ NO ₁₇	+	917	174	60	37	773	29	50	3.84
	Carbamazepine (CBZ)	298-46-4	236.27	C ₁₅ H ₁₂ N ₂ O	+	237.1	194	28	19	179	39	50	3.85
Nourolontia	Epoxycarbamazepine (Ep-CBZ)	36507-30-9	252.27	$C_{15}H_{12}N_2O_2$	+	253.3	179.9	28	28	236	12	50	3.2
Neuroleptic	Oxazepam (OZP)	604-75-1	286.71	$C_{15}H_{11}CIN_2O_2$	+	287.4	241	34	20	269.1	14	50	4.08
	Oxazepam (Glu-OZP)	6801-81-6	462.84	$C_{21}H_{19}CIN_2O_8$	+	463.2	287.1	26	15	269	26	15	3.34
	Diclofenac (DICLO)	15307-79-6	294.14	$C_{14}H_{11}CI_2NO_2$	+	296.1	250	22	10	214.1	25	100	5.5
	Ibuprofen (IBU)	15687-27-1	206.28	C ₁₃ H ₁₈ O ₂	-	205	161	17	7			50	4.06
Non-steroidal anti-	Hydroxyibuprofen (OH-IBU)	51146-55-5	222.28	C ₁₃ H ₁₈ O ₃	1	221.2	177	19	9	158.7	13	50	1.2
	Ketoprofen (KETO)	22071-15-4	254.28	$C_{16}H_{14}O_3$	+	255	209	29	12	105	22	100	4.14
	Salicylic acid (SCA)	69-72-7	138.12	$C_7H_6O_3$	-	137	92.6	30	14	64.7	28	70	1.16
	Acetaminophen (PARA)	103-90-2	151.16	C ₈ H ₉ NO ₂	+	152	110	25	15	90	10	50	1.24
Missellaneous	Acetaminophen Glucuronide (Glu-PARA)	16110-10-4	327.29	C ₁₄ H ₁₇ NO ₈	+	350	173.8	33	15				1.64
Miscellaneous	Caffeine (CAF)	58-08-2	194.19	C ₈ H ₁₀ N ₄ O ₂	+	195.1	137.7	37	18	109.7	22	50	1.35
	Hydrochlorothiazide (HCTZ)	58-93-5	297.74	C ₇ H ₈ ClN ₃ O ₄ S ₂	-	296.2	77.6	42	28	204.8	22	50	1.5
	Ethyinylestradiol (EE)	57-63-6	296.4	C ₂₀ H ₂₄ O ₂	-	295.2	144.9	54	40	183	35	50	4.07
Horresse	17β-Estradiol (βE)	50-28-2	272.38	C ₁₈ H ₂₄ O ₂	-	271.1	145	50	38	183	41	70	3.89
Hormone	Estrone (EO)	53-16-7	270.37	C ₁₈ H ₂₂ O ₂	-	269.1	145	53	35	183	36	70	4.14
	Progesterone (PGT)	57-83-0	314.46	$C_{21}H_{30}O_2$	+	315.2	97	32	24	109	26	50	5.77

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132

133 **3. Results and discussion**

134 3.1. Mass spectrometry optimization

The selection of optimum detection parameters (collision energy, cone voltage, ionization 135 mode) for each targeted compound was carried out by introducing a standard diluted single 136 solute solution at 5 mg.L⁻¹ directly in the mass spectrometer (without separation). The 137 pseudo-molecular ion ([M+H]⁺ or [M-H]⁻) was selected as the parent ion. Acetaminophen-138 glucuronide was ionized as sodium adducts ([M+Na]⁺) and the daughter ion correspond to 139 the sodium adduct of paracetamol obtained by the loss of glucuronic acid. Similar 140 fragmentation pattern with loss of carbohydrate group was observed with Glu-OZP ([M+H]⁺ 141 142 \rightarrow [M-Glu+H]⁺). In some cases, the standard molecules were purchased as sodium or 143 chloride salt so molecular weight of the commercial product indicated in the table 1 does not correspond to the formula of active compounds. So the molecular weights indicated in 144 the table 1 do not correspond to the mass of the pseudo molecular ion (AML, LOS, NAF, PRA, 145 TRI, DOX, ERY, LINCO and TYL). Positive mode was selected for most of the molecules and 8 146 analytes were ionized under negative mode because of their tendancy to loose a proton. 147 148 Two transitions are chosen for quantification and confirmation. If possible transition corresponding to the loss of simples fragments (ie. $-H_2O$ or $-CO_2$) has been prefered for 149 quantification or confirmation transition. Only one transition could be found to 4 molecules: 150 151 Ibuprofen, Gemfibrozil, Tamoxifen and Hydroxy-Tamoxifen. The results are presented in table 1. 152

153 3.2. On-line SPE method development

The efficiency of the SPE step was studied using two different types of SPE cartridge phases :
Oasis HLB (Direct Connect HP 20μm, 2.1x30mm) and XBridge C18 (Direct Connect HP 10μm,
2.1x30mm). The low energie interactions are predominant with the C18 phases, unlike for
HLB phases where the dipole-dipole interactions are brought into play. Table 2 presents
characteristics (log(Kow), pka, coefficient of dissociation, dipolar moment) of molecules. The
extraction yield was then calculated according to the following equation:

Extraction yield (%) = $100 * \frac{\text{Area}_{\text{SPE mode}}}{\text{Area}_{\text{conventional mode}}}$

For each compounds, the area obtained with the injection of 5mL of solution at 100 ng.L⁻¹ in SPE mode was compared to the area obtain in conventional mode (Vinj=5 μ L; C=100 μ g.L⁻¹).

162 The results are presented in figure 1. In a global overview the extraction yields are better with the Oasis HLB phase in comparison to the C18 phase. 11 molecules have slightly better 163 extraction yields with the XBridge C18 media. Given these results, Oasis HLB phase was 164 chosen for the SPE cartridges. The extraction yields are between 24% and 96%. Six 165 molecules, among them three hormones (ATE, TRI, DOX, EE, BE and EO) have extraction 166 yields inferior or equal to 50% but the signal is sufficient for our analysis given the 167 168 reproducibility of the extraction step. The loading time and flow rate influence the analyte retention onto the preconcentration cartridge. If the loading time is too short, a part of the 169 molecules of interest will not be collected in the cartridge. MeOH is used for the cartridge 170 conditioning during 3 minutes and UPW for the loading sample during 5.5 minutes at 171 172 2mL/min. 5mL of sample are injected onto the cartridge. Elution of our compounds is made 173 using the initial chromatographic conditions. The preconcentration method takes 8.5 174 minutes. The pH of samples and eluents was also optimized to try to improve the extraction 175 yields. The figure 2 shows the effect of pH (3, 7 and 9) on molecule's recovery yields. Most of

the targeted compounds were efficiently extracted at neutral pH values. The recovery yields 176 of thirteen molecules (LOS, GEM, TAM, OH-TAM, IFO, TYL, DICLO, PARA, CAF, CBZ, OZP, PGT 177 and ERY) do not show significant pH dependence. ATE, NAF and LINCO were comparatively 178 more recovered under neutral condition due to the amine/ammonium repartition for the 179 low pH values. DANO and OFX are amphoteric molecules and exhibit higher recovery yields 180 181 under acid extraction than under neutral conditions. AML and OFX have extraction yields superior to 100%, the differences may be included within the experimental errors. Three 182 183 hormones have a better extraction yields at basic pH while below 23% for an acid pH. The SPE appears globally controlled by the carboxylic functions. The best compromise to our 184 analytical method is the neutral pH. 185

Tableau 2: log(Kow), pka, coefficient of dissociation and dipolar moment of molecules

Molecule	Log(Kow)	pka	coefficient of dissociation	dipolar moment		
AML	3	8.6	5.00 10 ⁻⁵			
ATE	0.16	9.6	1.50 10 ⁻⁵	5.71		
LOS	1.19	5,5	8.80 10 ⁻³			
NAF	4.56	8.7	4.70 10 ⁻⁵	2.83		
PRA	1.35	4.5	5.60 10 ⁻³			
PRO	3.48	9.5	1.70 10 ⁻⁵			
GEM	4.77	4.7	4.40 10 ⁻³			
TRI	1.04	4.3/8.9	7.00 10 ⁻³			
ТАМ	3.24	8.76	4.20 10 ⁻⁵			
OH-TAM	4.74	3.2/6.4	6.30 10 ⁻⁴			
IFO 0.86 13.2		2.50 10 ⁻⁷				
DOX 2,37 3.5/7.7		3.5/7.7	$1.40 \ 10^{-4}$			
ERY	3,02	8.8	3.90 10 ⁻⁵			
OFX	0.65	6.1	9.40 10 ⁻⁴	7.2		
SUL	0.79	5.7	1.40 10 ⁻³			
TRP	0.91	7.1	2.80 10 ⁻⁴			
DANO	0,44	6.0	9.90 10 ⁻⁴			
LINCO	0,56	7.6	1.60 10 ⁻⁴			
SFZ	0.19	7	3.20 10 ⁻⁴	7.34		
TYL	1.63	7.7	$1.40 \ 10^{-4}$			
CBZ	2,77	7	1.00 10 ⁻⁷	3.66		

Ep-CBZ	1.58	15.9	$1.00 \ 10^{-8}$	
OZP	2,24	1.7/11.6	1.30 10 ⁻¹	
DICLO	4,51	4	8.00 10 ⁻³	4.55
IBU	3,79	4.5	5.30 10 ⁻³	4.95
OH-IBU	3,97	4.8	3.90 10 ⁻³	
КЕТО	3.12	4.45	6.00 10 ⁻³	
SCA	1,19	3	3.10 10 ⁻²	
PARA	0,49	9.5	1.80 10 ⁻⁵	4.55
CAF	-0.091	14	2.10 10 ⁻¹	3.71
HCTZ	-0,07	7.9	$1.00 \ 10^{-4}$	
EE	3,67	10.3	$7.00 \ 10^{-6}$	
βE	3.57	10.71	4.40 10 ⁻⁶	1.56
EO	3.69	10.4	6.00 10 ⁻⁶	3.45
PGT	4	18.9	3.50 10 ⁻¹⁰	





molecules in neutral pH





Figure 2: Extraction yields calculated for the 3 pH (3, 7 and 9) for all analytes

194 *3.3. Chromatographic conditions*

Three chromatographic columns packed with different stationary phases were studied, two 195 196 using the reversed phase mode: Acquity BEH C18 (100 mm x 2.1 mm ID, 1.7 µm) and Acquity 197 HSST3 (100 mm x 2.1 mm ID, 1.7 μ m). These two columns have the same stationary phase 198 but Acquity HSST3 should allow for better separation of polar molecules due to the greater 199 proportion of residual silanol groups. The third column has a polar stationary phase: BEH 200 amide (100 mm x 2.1 mm ID, 1.7 μ m) in order to separate the analyte using hydrophilic 201 interaction liquid chromatography (HILIC). Comparing the chromatograms obtained for the 202 C18 and HSST3 column, the results are guite similar. Seven minutes are required to obtain 203 sufficient separation. It should be underlined that the resolution between two consecutive 204 peaks was quite low. However, because the quantification was done using different MRM channels this poor resolution does not affects the analytical performances. 205

Figure 3 summarizes the results by plotting the polarity (log Kow) as function of the capacity factor of the molecule, molecules with k'<1 form the unretained groups with no log(kow) dependances. For the others, correlation between k' and log(kow) shows two adverse 209 behaviours in relation with the different stationary phase, BEH and HSST3 on the one part and HILIC on the second part. Reversed phase HPLC columns (BEH C18 and HSST3) provide a 210 satisfactory separation with k' ranging from 0.93 to 9.91 according to the polarity of the 211 considered compounds. However numerous analytes exhibit a high polarity and were poorly 212 retained using reversed-phase HPLC. Normal phase HPLC column (BEH Amide) provides 213 214 separation with k' ranging from 0.1 to 9.6. Molecules retained by the reversed phase HPLC column are not retained in normal phase HPLC with k'<1. Moreover, peak tailing are 215 observed for some molecules with HSST3 (SUL, GEM, DOX) and with HILIC column (PARA, 216 DANO, HCTZ, TRI). The best compromise for our analyses is to use the BEH C18 column. 217





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The mobile phase flow rate was 0.4mL.min⁻¹, corresponding to the optimum zone of the Van Deemter curve with this column [26]. The elution conditions were optimized. Two chromatographic separation methods were needed to quantify all the target analytes. Indeed, analytes with ESI+ detection have better sensitivity with acidified eluents (with 0.1% of formic acid) unlike molecules with ESI- detection which have better sensitivity with neutral eluents. Moreover, the combination of both positive and negative ionization mode during the same run does lead to a decrease of the sensibility.

The elution conditions start with 20% ACN/80% UPW during 1 minute followed by a gradient 90% ACN within 6 minutes and remain constant for 1 min before returning to initial conditions, details of the method are presented in Supporting information (Section B –

233 Figures S1-S3)

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Examples of chromatograms obtained with a solution of 50 ng.L⁻¹ in UPW and the eluent program are presented in Figure 4. 12 molecules elute within two minutes for the ESI+/acid eluent method. As mentioned above, the detection mode (MRM) allows an accurate quantification even if the resolution is low.





Figure 4: Chromatogram obtained at 50ng/L in UPW. a. first method with ESI+. b. Second method with ESI-241

242 3.4. Quantification limit and matrix effect

Standard addition method was selected for calibration method in order to minimize or 243 eliminated matrix effects. Figures 5 present examples of calibration curve for CBZ in UPW, 244 245 Groundwater (GW), Drinking water (DW) and Surface water (SW). Limit of quantification 246 (LOQ) were determined for all targeted compounds in UPW and GW with the equation given in figure 5a, in accordance with the AFNOR NF-T-90-210 norm for all analytes. GW could be 247 considered free of pharmaceuticals residues because GW is drawn from a well recovering 248 249 the waters on a small watershed without collective or on-site sanitation water release, and 250 UPW can be considered as a matrix blank. Negatively ionized molecules (EO, BE, EE, HCTZ, SCA, IBU, OH-IBU, GEM, PRA) have higher limits of quantification because the background 251 noise is more important than for ESI⁺. The values of the quantification limit of targeted 252 compounds are presented in figure 6a. LOQ values obtained range from 5 to 17ng/L. These 253 limits of quantification are sufficient for our purpose. 254

255 Measurement errors were incorporated by defining the 90% confidence intervals (figure 5b). Figures 5c and d show standard addition calibration lines of CBZ in GW and DW. 256 Comparisons of the slopes obtained with real waters to the slope obtain in the blank 257 (a_{GW}/a_{UPW}) and a_{DW}/a_{UPW} allow a comprehensive approach of the matrix effects. These slope 258 ratios are presented in figure 6b for all analytes. The matrix effect is a classical phenomenon 259 260 which can be very important in liquid chromatography coupled with mass spectrometry because of the ionization process may be drastically influenced by the presence of 261 interfering species. Many studies have already described this phenomenon especially with 262 wastewaters. The presence of organic or inorganic substance can cause inhibition (<1) or 263 enhancement (>1) of a compound's signal [27-29]. In our case, natural organic matter may 264 265 disturb the SPE step or mass ionization so the rationalization of the slopes provides a global overview of matrix effect but do not allow to identify the critical step. 266

In figure 6b, matrix effects are not significant when the ratio is close to 1. In drinking water
this ratio was close to 1 for most of the analytes, only AML has a ratio superior to 5.



Figure 5: a. equation of LOQ determanation. b. Exemple of standard addition for CBZ with 90% confidence interval. c.
 and d. Exemple of standard addition in GW and DW for CBZ

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The developed method was used to determine the concentration of 37 pharmaceuticals substances in inflow and outflow waters of 8 drinking water treatment plants (DWTP) in west of France. The samples were collected once a month between october 2013 and april 2015, resulting in an average of 100 inflow and 100 outflow concentration values for each molecule. Nine pharmaceuticals have not been detected or with concentrations below the LOQ (AML, TAM, OH-TAM, IFO, ERY, LINCO, EE, βE and PGT). Figure 7 shows the

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285 concentrations of 27 pharmaceuticals or metabolites in surface water as a box plot; this statistical representation summarizes the data, for each compound, by the mean values, 286 287 median value, first and third guartiles and observed extrema. 7 molecules (PARA-GLU, KETO, OH-IBU, DANO, PARA, SCA, CAF) have a mean concentration greater than 50 ng.L-1. 10 288 molecules were quantifieded with mean concentrations higher than 10 ng.L⁻¹ (GEM, CBZ, 289 290 DICLO, OZP, OFX, IBU, HCTZ, ATE, PRO and DOX). The last detected 10 molecules exhibit mean concentration lower than 10 ng.L-1 (SFZ, SUL, TRI, PRA, Ep-CBZ, TRP, EO, NAF, TYL, 291 292 LOS). For some molecules, large differences between the extrema are observed (PARA-Glu, KETO, OH-IBU, SCA). These differences depend on the sampling date essentially. It should be 293 underlined that median values are close to mean values indicating that extrema values do 294 not play an important role. The maximum observed concentration in surface water was 650 295 ng.L⁻¹ for KETO. Detection frequencies depend on compounds and range from 100% 296 occurrence for CAF and PARA and 9% for TYL. 13 molecules (PARA-Glu, KETO, OH-IBU, 297 DANO, PARA, CAF, SCA, DICLO, GEM, CBZ, OZP, OFX and ATE) were quantified in more than 298 299 50% of surface water samples. In drinking water (figure 8), six molecules (KETO, PARA-Glu, OH-IBU, DANO, PARA and CAF) were quantified in 90% or more of the drinking water 300 301 samples. These 6 molecules were also the most quantified molecules in surface water. The 302 overall mean concentration values are between 4 (OZP) and 327 ng/L. The maximum 303 concentration found was 650 ng/L for KETO. For drinking water, the same remark than for surface water may be made concerning the gap between minimum and maximum 304 305 concentrations: the eight drinking water treatment plants operate different treatment 306 chains with different type of water resources.



Figure 7: overall mean concentrations (\$), median value, first and third quartiles and extrema of 27
 molecules detected on average above LOQ in surface waters and detection frequencies (%, broken line).







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Figure 8: overall mean concentrations(�), median value, first and third quartiles and extrema for 14 molecules detected in tap waters and detection frequencies (%, broken line).

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4. Conclusion

A multiresidue analysis was developed using on-line solid phase extraction connected to 317 liquid chromatography coupled with tandem mass spectrometry in order to quantify residue 318 trace levels 35 pharmaceuticals compounds in surface and drinking water. The short 319 320 implementation time needed to achieve the preconcentration and the analysis, 17 minutes for the positive mode method and 15 minutes for the negative mode method is among the 321 most significant advantages of this method compared to off-line solid phase extraction. The 322 developed method with a preconcentration factor of one thousand showed detection limits 323 compatible with the study of environmental matrices with very low analyte concentrations. 324 325 The limits of detection and quantification are between 1.5 and 4 ng/L and 4 and 17ng /L, 326 respectively. Standard addition was chosen for the quantification of molecules in water samples to overcome the matrix effects and provide an accurate determination of targeted 327 compounds. Among all studied substances, doxicycline appeared to be the most affected by 328 329 a matrix effect. The developed methods were applied to eight surfaces and drinking water. 330 In surface water, 12 molecules could be quantified in almost all analyzed samples with a 331 maximum concentration value of 650ng/L for Ketoprofen. In drinking water, 5 molecules 332 could be regularly detected, with overall mean concentration values between 20 à 120ng/L.

333 References:

[1] Jorgensen, S.E., Halling-Sorensen, B., Drugs in the environment *Chemosphere 40*, 2000,
691.

[2] Gros, M., Rodriguez-Mozaz, S., Barceló, D. (2012). Fast and comprehensive multi-residue
analysis of a broad range of human and veterinary pharmaceuticals and some of their
metabolites in surface and treated waters by ultra-high-performance liquid chromatography

- coupled to quadrupole-linear ion trap tandem mass spectrometry. *Journal of Chromatography A*, 1248, 104–121.
- [3] Togola, A. and H. Budzinski (2008). Multi-residue analysis of pharmaceutical compounds
 in aqueous samples. *Journal of Chromatography A* 1177(1), 150-158.
- [4] Viglio, L., Aboulfald, K., Deneshvar, A., Prevost, M., Sauve, S. (2008). On-line solid phase
 extraction and liquid chromatography/tandem mass spectrometry to quantify
 pharmaceuticals, pesticides and some metabolites in wastewaters, drinking, and surface
 waters. *J. Environ. Monit.*, 10, 482–489.
- [5] Kim, S.D., Cho, J., Kim, I.S., Vanderford, B.J., Snyder, S.A. (2007). Occurrence and removal
- of pharmaceuticals and endocrine disruptors in South Korean surface, drinking, and waste
 waters. *Water Research*, 41, 1013-1021.
- [6] Miege, C., Choubert, J.M., Ribeiro, L., Eusebe, M., Coquery, M. (2009). Fate of
 pharmaceuticals and personal care products in wastewater treatment plants Conception of
 a database and first results. *Environmental Pollution*, 157, 1721–1726.
- 353 [7] Moldovan Z. (2006) Occurrences of pharmaceutical and personal care products as 354 micropollutants in rivers from Romania. *Chemosphere* 64,1808-1817
- [8] Mompelat, S., Le Bot, B., Thomas, O. (2009). Occurrence and fate of pharmaceutical
 products and by-products, from resource to drinking water. *Environment International*, 5,
 803-814
- 358 [9] Stackelberg, P.E., Gibs, J., Furlong, E.T., Mayer, M.T., Zaugg, S.D., Lippincott, R. (2007).
 359 Efficiency of conventional drinking-water-treatment processes in removal of
 360 pharmaceuticals and other organic compounds. *Science of the Total Environment*, 377, 255361 272.

362 [10] de Graaff, M.S., Vieno, N.M., Kujawa-Roeleveld, K., Zeeman, G., Temmink, H., Buisman,

363 C.J.N., (2011). Fate of hormones and pharmaceuticals during combined anaerobic treatment

364 and nitrogen removal by partial nitritation-anammox in vacuum collected black water.

365 *Water Research*, *45* (1), 375.

[11] Daughton, C.G., Ternes, T.A., (1999), Pharmaceuticals and personal care products in the
environment: agents of subtle change? *Environ. Health Perspect.* 107, 907.

[12] Vieno, N.M., Tuhkanen, T., Kronberg, L. (2006). Analysis of neutral and basic
pharmaceuticals in sewage treatment plants and in recipient rivers using solid phase
extraction and liquid chromatography–tandem mass spectrometry detection. *Journal of Chromatography A*, 1134, 101-111

[13] Viglio, L., Aboulfald, K., Deneshvar, A., Prevost, M., Sauve, S. (2008). On-line solid phase
extraction and liquid chromatography/tandem mass spectrometry to quantify
pharmaceuticals, pesticides and some metabolites in wastewaters, drinking, and surface
waters. *J. Environ. Monit.*, 10, 482–489.

[14] Carlsson, C., Johansson, A.K., Alvan, G., Bergman, K., Kuhler, T., (2006) Are
pharmaceuticals potent environmental pollutants? Part I: Environmental risk assessments of
selected active pharmaceutical ingredients. *Science of the Total Environment* 364 (2006) 67 –
87.

[15] Cleuvers, M. (2003) Aquatic ecotoxicity of pharmaceuticals including the assessment of
 combination effects. *Toxicology Letters* 142, 185-194.

[16] Gerrity, G., Gamage, S., Holady, S.J., Mawhinney, D.B., Quinones, O., Trenholma, R.A.,
Snyder, S.A., (2011). Pilot-scale evaluation of ozone and biological activated carbon for trace
organic contaminant mitigation and disinfection, *water research*, 45, 2155-2165

[17] Rodriguez-Mozaz, S., Lopez de Alda, M.J., Barcelo, D., (2007). Advantages and
 limitations of on-line solid phase extraction coupled to liquid chromatography-mass
 spectrometry technologies versus biosensors for monitoring of emerging contaminants in
 water. *Journal of chromatography A.*, 1152, 97-115.

[18] Trenholm, R., Vanderford, B., Snyder, S., (2009). On-line solid phase extraction LC MS/MS analysis of pharmaceutical indicators in water: A green alternative to conventional
 methods. *Talenta*, 79, 1425-1432.

[19] Miège, C., Favier, M., Brosse, C., Canler, J.P., Coquery, M., (2006). Occurrence of
betablockers in effluents of wastewater treatment plants from the Lyon area (France) and
risk assessment for the downstream rivers. *Talenta* 70, 739-744.

[20] Postigo, C., Alda, M.J.L.D., Barcelo, D., (2008) Fully Automated determination in the low
 nanogram per liter level of defferent classes of drugs and abuse in sewage water by On-line
 solid-phase extraction-liquid-chromatography-electrospray-tandem mass spectrometry,
 Anal. Chem., 80, 3123-3134.

[21] Viglio, L., Aboulfadl, K., Mahvelat, A.D., Prevost, M., Sauve, S., (2008). On-line solid
phase extraction and liquid chromatography/tandem mass spectrometry to quantify
pharmaceuticals, pesticides and some metabolites in wastewaters, drinking, and surface
waters, J. Environ. Monit. 10, 482-489.

[22] Alnouti, Y., Srinivasan, K., Wadell, D., Bi, H., Kavetshaia, O., Gusev, A.I., (2005).
Development and application of a new on-line SPE system combined with LC-MS/MS
detection for high throughput direct analysis of pharmaceutical compounds in plasma, *Journal of Chromatography A.* 1080, 99-106.

407 [23] Wang, S., Huang, W., Fang, G., He, J., Zhang, Y., (2008). On-line coupling of solid-phase
408 extraction ti high-performance liquid chromatography for determination of eostrogens in
409 environment, *Anal. Chim. Acta* 606, 194-201.

[24] Feitosa-Felizzola, J., Temime, B., Chiron, S., (2007). Evaluating on-line solid-phase
extraction coupled to liquid chromatography ion trap mass spectrometry for reliable
quantification and confirmation of several classes of antibiotics in urban wastewaters, *Journal of Chromatography. A.* 1164, 95-104.

414 [25] Van De Steene, J., Lambert, W., Comparison of matrix effects in HPLC-MS/MS and UPLC-

415 MS/MS analysis of nine basic pharmaceuticals in surface waters. *Journal of The American*

416 *Society for Mass Spectrometry* 2008, *19* (5), 713.

417 [26] Swartz, M., Murphy, B. (2007). L'Ultra Performance LC: l'avenir de la chromatographie

418 liquid. *Les Technologie de laboratoire,* n°3 Mars-Avril, 20-22

419 [27] Cimetiere, N., Soutrel, I., Lemasle, M., Laplanche, A., Crocq, A. (2013) *Environmental*420 *Technology*, 34(22), 3031-3041

421 [28] Kasprzyk-Hordern, B., Dinsdale, R.M., Guwy, A.J., (2008). The effect of signal 422 suppression and mobile phase composition on the simultaneous analysis of multiple classes 423 of acidic/neutral pharmaceuticals and personal care products in surface water by solid-phase 424 extraction and ultra-performance liquid chromatography–negative electrospray tandem 425 mass spectrometry. *Talanta*, 74, 1299–1312

- 426 [29] Balakrishnan, V.K., Terry, K.A, Toito, J., (2006). Determination of sulfonamide antibiotics
- 427 in wastewater: A comparison of solid phase microextraction and solid phase extraction
- 428 methods. *Journal of Chromatography A*, 113, 1–10.